

Assessment and Synthetization of Extension Needs Encompassing the Economic and Ecological Impacts of Wild Pigs Among Young Pine Plantations

by

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Abstract

The southeastern United States has recently seen its first increases in Longleaf pine (*Pinus palustris* Mill.) acreage after more than four centuries of decline thanks to ongoing restoration efforts. Connecting existing longleaf landscapes across this species' native range depends largely on the successful establishment of plantations on non-industrial private lands. As in Alabama, the majority of forest lands in southeastern states are owned by non-industrial private landowners. Seedling depredation by wild hogs (*Sus scrofa*) poses a threat to longleaf restoration efforts especially in young forest plantations. Improving the knowledge base about wild hogs among young forest plantations will help guide better management decisions for non-industrial private landowners and resource professionals. This dissertation addresses extensions needs for information regarding wild hogs and explores the relationship between wild hogs and young forest plantations from two perspectives.

First, an extension publication was created to synthesize and organize resource material pertaining to wild hogs in order to supplement the nine year gap since the last published bibliography. Second, a mail survey was conducted among non-industrial private landowners in Alabama to gain an understanding about wild hog damage and control in forest plantations. Findings were used to further explore the economics of wild hog damage and control for a stand level model utilizing longleaf. Third, a field study

was performed to determine wild hog's preference among seedling species and to observe ecological factors influencing seedling depredation.

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CHAPTER 1

Introduction

The longleaf pine (*Pinus palustris* Mill.) species has endured numerous anthropogenic hardships over the past four centuries, the last century alone has proved the most taxing. Irresponsible logging practices, tapping for pitch, fire suppression, the spread of agriculture, and the introduction of wild hogs are a few factors contributing to the major reduction of a once numerous resource (Frost 1993). During the first half of the 20th century, concern escalated over why longleaf regeneration was not succeeding. Along with fire suppression, wild hogs (*Sus scrofa*) were one of the main factors driving regeneration failure in longleaf pine forests (Frost 1993, Hopkins 1947, Wakeley 1954).

Present in over 44 states, wild hogs are receiving more attention than ever as the problems stemming from their presence become more evident (Mayer 2009a). The deleterious effects of wild hogs on longleaf regeneration both directly and indirectly result from a number of the animals' daily activities. The result of rooting activities can often lead to complete crop failure in forest operations (Campbell and Long 2009, Mayer 2009b, Wakeley 1954). Rooting can also indirectly affect longleaf regeneration by altering the vegetation structure.

Longleaf pine restoration efforts have the potential to be affected by growing wild hog population which currently has proved difficult to manage. This review of the history, ecology, and related research regarding longleaf pine and wild hogs will help bring understanding to an issue centuries in the making.

History and Ecology of Longleaf Pine

The historical range of virgin longleaf forests dominated a belt of land along southeastern states stretching from Virginia to Texas (Schwarz 1907). This 92 million acre expanse of forest was primarily dominated by longleaf pine (Frost 1993). Over thousands of years the landscape afforded ideal growing conditions for this species alone due to the frequent fire interval sparked by lightning (Frost 1993). In addition to lightning, Native Americans started fires in these forests to drive game species towards hunters (Frost 1993).

The understory of historic longleaf forests were comprised primarily of various types of grasses; the most common of which were wire-grass (*Aristida stricta* Michx.) and broomsedge bluestem (*Andropogon virginicus* L.) (Schwarz 1907). One unique aspect of longleaf pine is the canopy structure which allows ample sunlight to reach the forest floor facilitating the growth of grasses (Means 2007). Along with longleaf, grasses were another common species able to grow on a landscape subject to such frequent fire return intervals. Grasses served as an important ignition source for the low intensity fires longleaf require for proper growing conditions (Landers et al. 1995). The lack of natural land barriers allowed fire to burn unimpeded across the Southeast (Chapman 1932). The savannah-like conditions resulting from the longleaf landscape exposed to frequent fire

created a unique ecosystem capable of sustaining a high diversity of species richness (Peet 2007, Means 2007). Nearly 900 plants species exist in association with longleaf pine ecosystems, along with 100 bird, 36 mammal, and 170 reptile and amphibian species (NRCS 2011). A comprehensive list of the flora and fauna associated with longleaf pine ecosystems has been described further by Means (2007). Over one-third of the bird species and 69 percent of the mammalian species forage on or near the ground, and rely on the ground cover associated with landscapes maintained by fire (Van Lear et al. 2005). The key to these ecosystems sustaining such a high diversity of species richness is the open-canopied grasslands maintained by frequent fire (Means 2007).

Chapman (1932) explained the many fire adaptations of longleaf allowing it to thrive in the Southeast. First, during the initial growth period longleaf seedlings concentrate growth to the root rather than the stem. Second, nutrients are stored in the roots allowing it to put out new needles after the preexisting needles are burned off. Third, the thick, long needles unique to longleaf may protect the scaly bud from damage by fire. Fourth, longleaf put on thick, fire resistant bark when it commences stem growth. Lastly, young saplings lack many branches discouraging crown fires. Fire was also critical for successful germination and establishment through the elimination of competition. Longleaf have large seeds with limited dispersal range and seedlings are slow growing, remaining in the stemless grass stage for years (Landers et al. 1995). Additionally, fire is critical in reducing damage from brown spot needle blight. Brown spot needle blight is caused by a fungus which grows on the underside of dead grasses and spreads onto the needles of the seedlings by rain splash (Greene 1931). The survival

of the saplings is unaffected by this disease as long as fire burns off the needles harboring the fungus.

It did not take long for the first settlers to recognize the valuable resource spread across the southeastern landscape. Roland Harper (1913), a field botanist of the early 20th century, described longleaf as having more uses than any other tree in the world. Frost (1993) documented the historic events which brought about the drastic reduction of longleaf forests. Prior to the 1700's, impacts to longleaf by early European settlers were limited to coastal regions along navigable waterways. Commercial logging began to have a more serious impact on the longleaf population when water-powered sawmills were constructed beginning in 1714. Over the next 50 years, hundreds of these mills materialized over the Southeast but were still limited to areas accessible by rivers.

The naval store industry is responsible for much of the disappearance of longleaf during the 19th century. The sap collected from pine trees was used to produce tar, pitch, rosin, and turpentine and collectively referred to as naval stores. Early records indicate naval stores existed in Virginia some 200 years before the industry peaked in North Carolina. The practice of boxing longleaf for crude turpentine left the trees damaged beyond repair. With the invention of the copper still in 1834, turpentine distillation was made vastly more efficient and profitable. The copper still allowed for the commercial exploitation of longleaf from North Carolina to Texas. Advances in steam technology and logging railroads during the period of 1870-1920 devastated the last remaining expanses of virgin forest in the South. By this point longleaf was being shipped around the world because of its value as a superior lumber product.

Frost (1993) postulated the failure of longleaf to regenerate itself was driven by three factors. First, around the turn of the 19th century, much of the southern landscape had been converted to open pastureland or agricultural fields. Longleaf, limited by the short range of seed dispersal, could not restock these open areas. Also, historic longleaf sites were being restocked with faster growing pine species like loblolly (*Pinus taeda*) and slash pine (*Pinus elliottii*). Second, there was a period during 1910-1930 where prescribed or intentional burning was outlawed due to the misconstrued belief that fire was harmful to southern forests. Fire suppression from the southern forest system proved detrimental to longleaf's ability to regenerate itself and allowed competition to take over the understory. Third, the open grazing of livestock was practiced up until the end of the 19th century. These free ranging livestock would have significantly impacted longleaf regeneration, especially the hog (Frost 1993).

Currently, the small numbers of longleaf forests resembling pre-settlement conditions are second-growth forests (Landers et al. 1995). Costa and DeLotelle (2007) describe most longleaf forests found scattered across the South as more even-aged structured and with denser understories composed primarily of shrub cover. According to Costa and DeLotelle (2007), this alteration of forest structure and loss of herbaceous cover has resulted in a reduction in faunal diversity and extirpation of many vertebrate species. Numerous species of plants and animals are now listed as endangered, threatened, and under conservation concern as a result of the disappearance of longleaf pine habitat (Means 2007, Van Lear et al. 2005). These at-risk species are described in more detail by Means (2007). Two endangered keystone species who have received the most public attention are the red-cockaded woodpecker (*Picoides borealis*) and the

gopher tortoise (*Gopherus polyphemus*) (Costa and DeLotelle 2007, Means 2007, Van Lear et al. 2005).

There are many who would regard the longleaf species as one of the most treasured forest resources in the southern United States (US) (Alavalapati et al. 2007). Longleaf restoration efforts have become considerably more commonplace due to a rising public preference for their valuable market outputs (timber and non-timber products) and non-market outputs (biodiversity, ecological services, and aesthetics) The US Forest Service, Department of the Interior, and US Department of Agriculture's (USDA) Natural Resources Conservation Service (NRCS) are highly involved in the restoration of longleaf pine to its historic range (Alavalapati et al. 2007, LPC 2017). More comprehensive details concerning the policies and cost-share programs created to incentivize landowners to plant more longleaf are described by Alavalapati et al. (2007). Private landowners play a critical role in the future of longleaf pine and with a unified effort perhaps this species will once again become a substantial component of southern forests.

History and Ecology of Wild Hogs

Mayer (2009a) composed the most complete historical account of wild hogs in the US to date. The origins of wild hogs are widespread and varied, for this reason it is nearly impossible to narrow down what species were introduced and when. There were essentially two types of *S. scrofa* introduced to the US, the Eurasian wild boar and domestic swine. It is commonly accepted among scientists and professionals that *S.*

scrofa encompasses the Eurasian wild boar, domestic swine, feral hogs, and any hybrids found in the US.

Spanish explorers Hernan Cortes and Hernado De Soto hold the first documented introduction of hogs to the continental US in the 1500's (Mayer 2009a). Hogs were an important food source to early explorers making them an essential commodity to bring on expeditions. Pigs were also a highly sought-after food source for Native Americans, who frequently traded or stole pigs from the Spaniards. Many other Spanish, French, and English explorers and colonists came after De Soto bringing with them more hogs. Native Americans would later provide colonist with salted pork and corn proving critical for their survival during their initial establishment. Free-ranging hogs proved to be the most economical way to raise pigs. Over the ensuing centuries open range practices of early settlers, farmers, and Native Americans would facilitate the establishment of feral populations (West et al. 2009).

It was during the 19th century that the Eurasian wild boar was introduced to the US by wealthy sportsmen (Mayer 2009a). Though the animals were kept in fenced hunting preserves, they would inevitably escape. The first of these hunting reserves was in New Hampshire, followed by more in North Carolina and Texas. These marked the most notable populations of escaped hogs who would later make up the populations currently seen in the US. The continued expansion of wild hogs across the country has been primarily manmade through the translocation by hunters favoring this game species. Interbreeding has occurred between the Eurasian boar and the feral pigs producing many of the hybrids currently found in the wild.

In addition to illegal translocation and release, the expansion of wild hog populations in North America is largely due to a prolific reproductive ability (Mayer 2009a). Once introduced to an area, the reproductive potential and adaptable biology of these animals enables them to quickly become established (Seward et al. 2004, West et al. 2009). Male wild hogs become sexually mature as young as 4-5 months and reach puberty within their first year (Comer and Mayer 2009). They are capable of breeding year round, though most breeding is done by larger, older males. Females are sexually mature as young as 3-4 months old and also reach puberty within their first year. The average litter size is 5-6 piglets and females can become pregnant within a month after giving birth (Comer and Mayer 2009, Graves 1984). Gestation periods last 112 to 120 days and females typically reproduce once a year, though they are capable of producing nearly two litters a year (Comer and Mayer 2009).

Most of the damage relating to wild hogs comes from aggressive foraging behaviors, specifically rooting (Mayer 2009b). Rooting often takes place in the winter or early spring when above-ground resources are scarce (Ballari and Barrios-Garcia 2014). These animals root to obtain food by breaking up and loosening soil in a search for roots, tubers, fungi, and burrowing animals (Mayer 2009b). Wild hogs are opportunistic omnivores (or generalist feeders), with food selection being influenced by energy requirements, food availability, and seasonal and geographic variations (Ballari and Barrios-Garcia 2014). Ballari and Barrios-Garcia (2014) synthesized scientific literature pertaining to wild hog diet and found a number of studies showed diet varied based on seasonally available foods. The study concluded that wild hogs consume above-ground vegetable parts in the spring, fruits year-round except the spring, and hard mast during

autumn and winter. Plant matter makes up 90 percent of wild hog's diet while invertebrates play a small but important role. Invertebrates are eaten throughout the year and research suggests they are a required food source for wild hogs.

As ecosystem engineers, wild hogs are capable of impacting forest communities on a number of different levels (Campbell and Long 2009, Crooks 2002). Many studies have examined how the activities of this species affect forest ecosystems by disturbing soils, mainly by rooting, ultimately changing the understory composition (Arrington et al. 1999, Bratton 1975, Chavarria et al. 2007, Engeman et al. 2003, Fagiani et al. 2014, Howe et al. 1976, Ickes et al. 2001, Lacki and Lancia 1986, Wirthner et al. 2012, Wood and Roark 1980). The majority of studies show rooting in introduced ranges has a negative effect on plant communities suggesting that plants are not adapted to wild hog disturbance (Ballari and Barrios-Garcia 2014). Other studies emphasized the wild hog's ability to reduce forest regeneration (Bruinderink and Hazebroek 1996, Hanson and Karstad 1959, Ickes et al. 2005, Siemann et al. 2009, Sweitzer and Van Vuren 2002).

Direct and Indirect Impacts of Wild Hogs on Longleaf Forests

Wild hogs have had a notable impact on forestry and timber resources (Campbell and Long 2009, Mayer 2009b, Mayer et al. 2000). The direct impacts most often involve pine species and include: girdling mature trees through rubbing, damaging the lateral roots by rooting or chewing, and damaging the bark by tusking (Mayer 2009b). Another direct impact of wild hogs, and the most economically costly, is their predation of planted pine seedlings (Mayer 2009b, Mayer et al. 2000). This primarily involves longleaf in its grass stage, but has also been observed with other southern pines like loblolly pine, slash

pine, and pitch pine (*P. rigida*) (Wakeley 1954). A single hog is reportedly capable of rooting up to six longleaf pine seedlings a minute, destroying an estimated 400-1000 seedlings a day (Hopkins 1947, Wakeley 1954).

Lipscomb (1989) showed wild hogs have a profound effect on longleaf pine regeneration. In South Carolina, he observed after two growing seasons, only eight of 845 longleaf pine seedlings survived hog predation in an unfenced area, versus 64 percent of seedlings surviving in fenced-in sites. Wood and Roark (1980) concluded the hogs are not actually consuming the pine saplings, but instead are chewing on the roots to access the sap and starches, then discarding the woody tissue. The damage done to the rootstock and lower portion of the stem result in seedling mortality and repeated over a large enough area can cause regeneration failure (Mayer 2009b).

Rooting indirectly affects longleaf regeneration by creating environments that favor pioneer and exotic species (Barrios-Garcia and Ballari 2012). Similarly, Singer et al. (1984) found rooting in deciduous forests can reduce as much as 80 percent of the understory and result in a change in vegetation composition. Siemann et al. (2009) found a significant increase in woody exotic species took place in areas where rooting occurred. Increasing the amount of exotic and pioneer species would cause longleaf seedlings to remain in the grass stage for a longer period of time and extend the period of vulnerability to wild hogs. Additionally, Singer et al. (1984) found wild hog activities can expose anywhere from 560-1130 tree roots per acre by rooting. Though many of the roots were not broken, they were exposed and vulnerable to the environment.

Wallowing often takes place in conjunction with rooting (Mayer 2009b). Though wallowing does not disturb as large of an area as rooting activities, it can indirectly cause mortality in longleaf seedlings by crushing, trampling, or pulling up trees within a wallowing pit. Wallow pits and rooted areas become fire barriers which can affect how efficiently a prescribed burn (or natural fire) carries through a stand.

There are many research topics of wild hogs in forest settings, but very few pertain to longleaf. Wild hogs affinity for longleaf pine seedlings has been documented over the past century but little research has been done other than showing that excluding wild hogs improves longleaf seedling survival (Lipscomb 1989). Additional research is needed in the future to better understand the connection between wild hogs and longleaf. Such an understanding could lead to better trapping and management techniques. Currently, the best option for protecting forest stands from wild hogs is to use a combination of available management techniques including exclusion fencing, general hunting, hunting with dogs, and trapping (Campbell and Long 2009). Managing wild hogs is helpful on a short term basis while longleaf seedlings are in the grass stage but an efficient long term wild hog management solution has yet to be found.

Dissertation Synopsis

This dissertation was created to meet research and extension needs concerning the impact of wild hogs on young forest plantations. Though there is a strong emphasis on wild hog damage to longleaf pine, the trends and topics discussed in the following chapters have broader implications. The research objectives defined below are a positive step towards increasing the knowledge base on wild hogs in the areas of economics,

ecology, and extension. The first research objective was to address the need for an updated bibliography of recent published resource material pertaining to wild hogs. The second research objective was to estimate and examine the economic impacts of wild hogs in non-industrial privately owned forests within Alabama. The third research objective was to determine wild hog's preference among seedling species and to observe ecological factors influencing seedling depredation. By achieving these research objectives, a greater understanding about wild hogs will contribute to improving the management of this species. To achieve the research objectives, this dissertation presents the results of three separate projects:

1. The first project (Chapter 2) built on the previous works of Wolf and Conover (2003) and Mayer and Shedrow (2007). This chapter synthesizes and organizes a number of different resource materials pertaining to wild hogs into a single searchable online document. Materials in the updated bibliography include selected journal articles, books, extension publications, conference proceeding articles, Internet publications, and academic theses. This bibliography was created to supplement the nine year gap since the last bibliography of works concerning wild hogs. Subject categories are included with each entry that best describe the nature of the material and how it applies to wild hogs. When available, the country where the work originated was included. Along with being included as the second chapter of this dissertation, the bibliography has been published by the Alabama Cooperative Extension System and is available at the following link:
<http://www.aces.edu/pubs/docs/F/FOR-2047/FOR-2047.pdf>.

2. The second project (Chapter 3) involved a survey sent to non-industrial private landowners across Alabama. The survey was designed to simultaneously capture information related to wild hog presence on private lands, damage to forested areas, forest types, control methods, and hunting. Descriptive statistics were used on the survey data and the results were organized by major physiographic region. Survey results revealed the level and cost of wild hog damage and control taking place in Alabama. This information was then used to build a stand level economic model of a longleaf pine stand in order to explore the potential economic implications of landowner responses to wild hog damage. Sensitivity analyses were performed on the costs of damage and control to observe how prices fluctuated when exposed to a range of potential discount rates and amounts of damage.
3. The third project (Chapter 4) involved a field study where wild hogs preference towards pine and hardwood seedlings were observed at two sites. Five tree species were planted at each site and observational visits were made monthly throughout the study period. One site received a heavy amount of hunting pressure while the other site did not. Site conditions were recorded so ecological inferences could be made about wild hog damage to planted seedlings.

CHAPTER 2

Wild Pig (*Sus scrofa*): An Update Bibliography

2.1. Introduction

This bibliography is a compilation of selected journal articles, books, extension publications, conference proceeding articles, Internet publications, and academic theses and dissertations from 2007 to present. This bibliography was created to supplement the nine year gap since the last bibliography of works concerning wild hogs published by John J. Mayer and C. Barry Shedrow “Annotated Bibliography of the Wild Pig (*Sus Scrofa*)” (2007). Information on wild hogs from works published before 2007 can be found in Wolf and Conover’s (2003), “Feral pigs and the environment: an annotated bibliography”. As with Mayer and Shedrow’s bibliography, the purpose of this document is to provide a readily accessible summary of literature on wild hogs. Only works pertaining specifically to wild hog research and management were selected for this bibliography. The author did not include entries which mentioned wild hogs in a general/minimal manner; nor did the author include references concerning all species in genus *Sus*. References comparing domestic pigs to their wild counterparts were included. In a few cases, general studies with wild hogs as a primary focus or sample group were included.

Due to the broad nature of work being done around the world regarding wild hogs, it should be noted that this bibliography does not include every article related to the species, but is merely a starting point for those interested in learning more about *Sus* species. Sources were identified using a computer database through Auburn University Libraries, and on the Internet's World Wide Web. The entries are followed by subject categories to assist the users of this bibliography to find the appropriate references they may need. Topics pertaining to subject categories are defined in Mayer and Shedrow's (2007) bibliography; those topics include, archaeological/paleontology, behavior, contamination, control/management, damage, diseases/parasites, domestication, ecology, economics, folklore/fiction, food habits, genetics, history, hunting, morphology, physiology, population biology, predation, radioecology, reproduction, and taxonomy. For most studies, the country where the study was conducted is noted in parenthesis after the subject categories associated with each entry. This bibliography will continue to be updated in the future. Please forward suggestions for the articles that were not included or published after 2016 to:

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2.2. Bibliography

A

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CHAPTER 3

Economic Estimates of Wild Hog (*Sus scrofa*) Damage and Control Among Young Forest Plantations in Alabama

3.1. Introduction

3.1.1. Current Status of Wild Hogs in the United States and Their Potential Impact on Alabama's Forests

Non-native wild hogs (*Sus scrofa*) rank as one of the top three most prolific invasive animals in the United States (US) (Pimentel 2007). Now present in over 44 states, the rapid expansion of wild hogs has been facilitated by illegal translocation and a naturally high fecundity rate (Mayer 2009a). Once introduced to an area, the reproductive potential and adaptable biology of these animals enables them to quickly become established (Seward et al. 2004, West et al. 2009). This species is particularly problematic to landowners because of their tendency to travel in groups and ability to cause extensive damage to timberlands, pastures, and agriculture crops (Graves 1984, Seward et al. 2004, West et al. 2009). Problems stemming from the presence of wild hogs are not new to the US; however, in recent years more attention is being given to the growing range of this species and the accumulating evidence of their deleterious impacts on agricultural and forested areas (Campbell and Long 2009, Sloodmaker et al. 2017).

Agriculture, forestry, and other related sectors are major contributors to Alabama's economy, with a combined value of well over \$100 billion (Fields et al. 2013). Wild hog damage has the potential to affect over 90 industrial sectors related to agriculture and forestry within the State. As the State's second largest manufacturing industry, forestry produced an estimated \$14.8 billion of products in 2013 (AFC 2016).

Ninety-four percent of Alabama's timberland acreage is privately owned with around 50 percent of the total timberland being pine plantations (AFC 2016). With Alabama's land area comprised of 27 percent farm operations and 69 percent timberlands, the deleterious effects from wild hog presence in the State is a major cause for concern (AFC 2016, USDA 2016). With isolated populations of wild hogs in at least 64 counties (Conley et al. 2014), this species has emerged as one of the leading wildlife problems. One wildlife biologist working for the Alabama Department of Conservation and Natural Resources describes wild hogs as "the most destructive nuisance animal ever brought to Alabama" (Jaworowski 2011).

With non-industrial private landowners in control of the majority of Alabama's agriculture and forest lands, the successful management of these operations is critical to the State's natural resource industry. In 2015, the Natural Resources Conservation Service (NRCS) allocated \$100,000 in financial assistance to Alabama landowners from select counties to address the growing threat from wild hogs. These funds were made available through the Environmental Quality Incentive Program (EQIP) to help cover the costs associated with wild hog control. Whether or not the monetary aid being offered is substantial enough to mitigate the damage being done to the State's agriculture and forest

lands has yet to be determined. It does, however, suggest the wild hog problem has grown serious enough in Alabama to warrant an assistance program for control efforts. The State currently has a no closed season/no bag limit hunting policy for wild hogs but efforts like these rarely make any substantial progress towards population reduction (Jerrolds et al. 2014, Massei et al. 2011).

Funding to manage wild hogs and protect valuable resources is finite and must be applied carefully in order to optimize its use. When economic restraints govern the management actions that can be taken towards controlling wild hogs, the metric for success of the management actions are measured by the amount and value of resources protected (Engeman et al. 2003). Therefore, providing an estimation of the monetary value to the damage wild hogs inflict on a resource would permit economic analyses to help guide and evaluate management actions (Engeman et al. 2003). Wild hog damage estimates often extend to agricultural crops (Anderson et al. 2016, Herrero et al. 2006, Schley and Roper 2003, Seward et al. 2004), while damage estimates among forest plantations are lacking.

3.1.2. Wild Hog Damage to Forestry and Timber Resources

The most comprehensive review of scientific literature describing wild hog damage to forests was done by Campbell and Long (2009). Many studies described forest ecosystem damage but few studies were mentioned which described damage to forest plantations while none provided economic estimates for the cost of damage. Most reports and scientific literature on wild hog predation of seedlings concerns natural regenerated pine (*Pinus* spp.), oak (*Quercus* spp.), and beech (*Fagus* spp.) species (Bruinderink and

Hazebrook 1996, Hanson and Karstad 1959, Ickes et al. 2005, Lipscomb 1989, Siemann et al. 2009, Sweitzer and Van Vuren 2002). Few reports or research exists which specifically address wild hog predation on planted seedlings. The only reports of wild hog predation of planted southern pines came after an intensive planting effort by the U.S. Forest Service across the Southeast in the mid-1900's (Conley 1977, Lucas 1977, Wakeley 1954). Entire plantations of longleaf pine (*Pinus palustris* Mill.) and slash pine (*Pinus elliotii*) were reported to have been destroyed by wild hogs in Alabama, Florida, Louisiana, and Mississippi (Lucas 1977, Wakeley 1954).

Mayer et al. (2000) is the only study of which we are aware that documents wild hog predation among planted hardwood seedlings. The group found that of nine hardwood species planted in a wetland restoration area in South Carolina, cherrybark oak (*Quercus pagodaefolia*), swamp chestnut oak (*Q. michauxii*), water hickory (*Carya aquatica*), and swamp tupelo (*Nyssa sylvatica* var. *biflora*) were the only species impacted by wild hog foraging activities. Non-affected seedling species included water oak (*Q. nigra*), green ash (*Fraxinus pennsylvanica*), persimmon (*Diospyros virginiana*), bald cypress (*Taxodium distichum*), and water tupelo (*Nyssa aquatica*). Though the study was informative about wild hog's preference among planted hardwood species, it lacked an economic component quantifying the cost of damage in the wetland restoration area.

Trees are most vulnerable to wild hogs during the initial years after planting or germination (Mayer 2009b, Sweeney et al. 2003). A single hog is reportedly capable of rooting up to six longleaf seedlings a minute, destroying an estimated 400-1000 seedlings a day (Hopkins 1947, Wakeley 1954). These animals have the potential to cause

complete crop failure in young timber plantations while seedlings are in their initial growth stages. In South Carolina, Lipscomb (1989) observed after two growing seasons, only eight of 845 longleaf seedlings survived hog predation in an unfenced area, versus 64 percent of seedlings surviving in fenced-in sites.

Wild hogs will occasionally cause damage to slash pine and loblolly pine (*Pinus taeda*), but the most extensive damage is reported to occur with longleaf (Wakeley 1954). Longleaf is unique among other planted tree species in that they have evolved with landscapes exposed to frequent fire. While other tree species focus energy into rapid vertical growth during initial stages of development, longleaf may remain in a fire-resistant grass stage for several years before initiating vertical growth (Croker and Boyer 1975). This development trait, along with a few other adaptations (long needles, thick bark, thickly scaled bud), allows longleaf to survive on landscapes with frequent fire intervals. During the grass stage, longleaf grows a thick tap root which may prove more appealing to wild hogs compared to root stems of other planted species (Hopkins 1947). Wood and Roark (1980) concluded the hogs are not actually consuming the pine saplings, but instead are chewing on the roots to access the sap and starches, then discarding the woody tissue.

The majority of longleaf forests in the US occur on non-industrial private lands (AFC 2016). Longleaf is desirable to landowners because it is generally more resistant to fire, hurricane damage, and pine bark beetle attacks than other southern pine species (Landers et al. 1995). Longleaf is also economically attractive because it produces a higher quality timber (North Carolina Forest Service 2011) and pine straw (Dyer 2012)

than loblolly and slash pine. On many sites longleaf can grow competitively or even exceed the growth of other southern pine species (Kush et al. 2007).

The structure of longleaf forests is ideal for multiple-use management because of the unique canopy structure which allows ample amount of sunlight to reach the forest floor (Franklin 2008). Alternative forestry (e.g., silvopasture, agroforestry, forest farming) has the potential to be more profitable than traditional forestry. Stainback and Alavalapati (2004) developed a stand level economic model of a silvopasture system utilizing longleaf to explore its potential profitability. The profitability of a silvopasture system was then compared with traditional pasture and traditional forestry scenarios. In addition, they also incorporated carbon payments and the effect of lengthening rotations to produce red-cockaded woodpecker (RCW) (*Picoides borealis*) habitat. Silvopasture was found to be more profitable than both traditional forestry and tradition pasture with and without carbon payments. The opportunity cost for increasing the rotation age to create RCW habitat was significantly less for silvopasture than for traditional forestry. The opportunity costs for both scenarios also decreased as carbon payments increased. The group found silvopasture to be an attractive land use opportunity for landowners wanting to plant longleaf on pastureland and take advantage of incentives for providing RCW habitat.

Longleaf forests have a high non-commodity value in terms of aesthetics and wildlife. Longleaf forests are among of the most diversified ecosystems in North American in terms of floral and faunal species (Means 2007, Peet 2007). These forests are good habitat for game-species (Franklin 2008) and can bring a landowner additional

income from hunting leases. There are also a number of non-game species endemic to longleaf pine ecosystems, two of which are endangered keystone species, the gopher tortoise (*Gopherus polyphemus*) and RCW (Means 2007). A number of studies have been done to explore if RCW habitat management can be profitable while producing timber (Lancia et al. 1989, Roise et al. 1990, Roise et al. 1991). Roise et al. (1991) developed a stand level economic model of a longleaf stand utilizing a shelterwood system with the addition of pine straw production. They concluded desirable RCW habitat can be profitably maintained for a longleaf operation given their management scenario. They also found harvesting pine straw contributed more income to the operations profitability than the timber revenue.

3.1.3. Research Needs and Project Description

Stand level economic models, like those described earlier in the studies by Stainback and Alavalapati (2004) and Roise et al. (1991), can be useful tools for evaluating the financial feasibility of growing timber. We are not aware of any studies which examine the economic impact of wild hog damage and control to young forest plantations, a concerning fact considering privately owned forests constitute nearly all of Alabama's timberland acreage and much of the forestland across the Southeast. Wild hogs can frustrate the efforts of landowners to establish forest plantations by incurring additional replanting costs after initial site preparation and planting expenses have been invested. In 2016, a landowner establishing a forest plantation would spend anywhere from \$29-141 per acre on site preparation, \$59-84 per acre on labor for planting, plus the additional cost of seedlings (which vary widely depending on supplier, species, and

number of trees planted per acre) (Maggard and Barlow 2017). With depredation of planted pine seedlings by wild hogs being an area of considerable economic costs when it comes to forest plantations (Mayer 2009b), any additional research pertaining to this topic could prove valuable to landowners and resource professionals.

The goal of this project was to fill in the knowledge gap concerning the economic impact of wild hogs to young forest plantations. A mail survey was conducted among non-industrial private landowners in Alabama to gain an understanding about wild hog damage and control in forest plantations. The survey was designed to simultaneously capture information related to wild hog presence on private lands, damage to forested areas, forest types, control methods, and hunting. When estimating the cost of wild hogs it is important to not only include the costs from the physical damage to the stand, but also the incurred costs from control measures aimed at preventing future damage. For this reason we included control costs as a part of our economic analyses. We then created a stand level economic model to explore potential outcomes of private forest landowner's responses to wild hog damage. Lastly, we utilized sensitivity analyses to examine how varying discount rates and amounts of damage might affect the cost of wild hog damage and control. Such information is beneficial in guiding forest management decisions as the threat from wild hogs becomes more widespread.

3.2. Methods

A number of tasks were created in order to guide this project through the research process. These tasks were designed to help accomplish the goal of providing information about the economic impact of wild hogs to forest stands. Five main tasks were associated with this portion of the study:

Task 1. Develop a questionnaire capable of soliciting information from private landowners regarding their forest stands and the presence of wild hogs on their property.

Task 2. Conduct a survey of non-industrial private landowners in Alabama to determine a reasonable estimate for the economic impact of wild hogs on young forest plantations.

Task 3. Perform descriptive statistics on survey response data and organize the results by physiographic region.

Task 4. Develop a stand level economic model and explore possible outcomes of landowner responses to wild hog damage.

Task 5. Using the stand level economic model, perform sensitivity analysis to determine how varying discount rates and levels of damage would affect the overall cost of damage from wild hogs.

3.2.1. Overview of Questionnaire Development

The questionnaire created to accomplish Task 1 was based on a prior survey conducted during the summer of 2015 (Anderson et al. 2016). In their survey, farmers from 11 states in the southern US were surveyed to assess the economic impact of wild hogs on crops, livestock, farmland, and the corresponding costs associated with wild hog control. Portions from the survey used for estimating damages were restructured to include questions relating to forestry. The survey was designed to simultaneously capture information related to wild hog presence on private lands and in forest stands, damage to forested areas, forest types, control methods, and hunting.

The questionnaire was designed to take 10 to 15 minutes to complete. Question types included multiple choice and fill-in-the-blank. Ample margin space was provided to give respondents the opportunity to write comments or include additional information. Questions were designed to minimize confusion, require minimal response effort, and maximize the amount of desirable information.

3.2.2. Timber Stand Section of Survey

Information on damage to young forest plantations was inferred by the questions listed in Appendix A. Landowners could choose to include their top two timber crops planted each year from 2013 to 2015. Even if the landowners did not experience wild hog damage, the survey questions were structured in such a way that useful information about their forest plantation could still be captured. Questions in this section also included

asking landowners to provide the number of acres damaged by wild hogs for each tree species planted and the percentage of seedlings destroyed on those acres.

Additional questions solicited landowners to report the frequency with which they had observed wild hogs from a number of general land cover types within their property. Participants were then asked to describe their property by assigning the percent of their land which best corresponded to a selection of predefined land cover categories (Appendix A)

3.2.3. Wild Hog Control Section of Survey

In this section of the questionnaire, landowners were asked to give information on the efforts used to control wild hogs on their property (Appendix A). Specifically, participants were questioned about the control methods used, the dollar amount invested in each method, and their perception on how effective the method was at controlling wild hogs. Questions about fencing, both non-electric and electric, had to be formatted differently from the control methods because they were categorized as fixed-cost. Lastly, landowners were asked to report how many wild hogs were killed on their land in 2015 and if they sought help from any county, state, or federal agencies in response to wild pig damage on their property.

3.2.4. Willingness-to-pay (WTP)/Willingness-to-accept (WTA) Section of Survey

In order to gauge landowner preferences for wild hog density and willingness to pay for eradication, participants were asked to select a category which best reflected their situation and opinion of wild hogs on their property or area (Appendix A). In responding

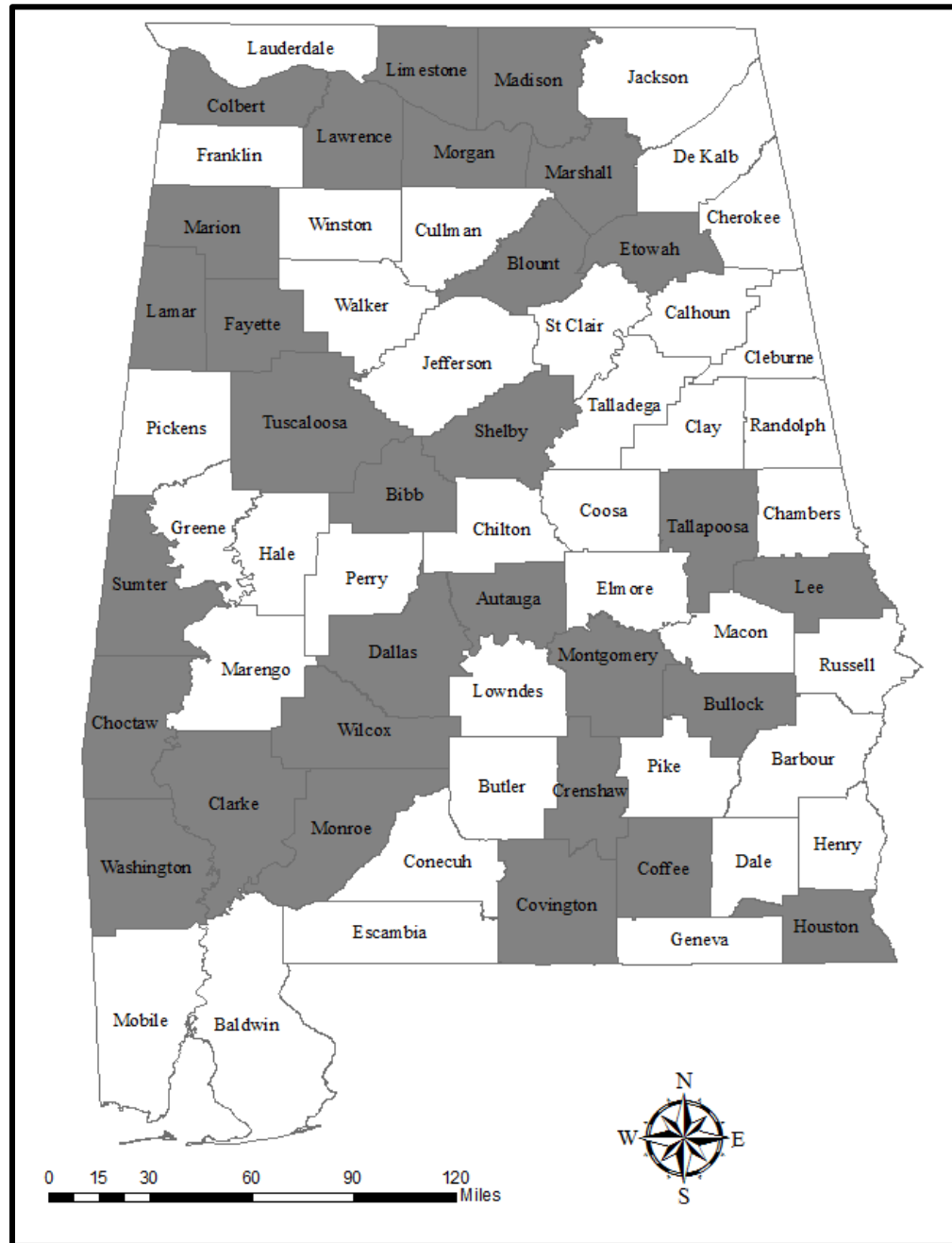
to this question, landowners would indicate the current status of wild hogs on their property and whether they would prefer more, fewer, or no change in wild hog density. The questionnaire was designed to group participants into those who preferred wild hogs and those who did not. Survey takers would then indicate a dollar amount from \$0 to \$1000 which they would be willing to pay or be paid annually on a per acre basis for eradication. Landowners who had wild hogs on their land but preferred fewer were grouped with those not wanting pigs because both associate negative values with wild hog presence. Other landowners, who desired to be supplemented for eradication because of the positive values they associated with wild hogs, were grouped together. This questionnaire assumes participants understood the hypothetical eradication would be permanent and that eradication costs would be shared regionally.

3.2.5. Survey Implementation

In Task 2, survey questionnaires were administered and coordinated by the Auburn University School of Forestry and Wildlife Sciences during the summer of 2016. Thirty of the 67 counties in Alabama were randomly selected to receive survey questionnaires (Figure 3.1). A stratified random sample of approximately 1,200 private landowners was taken from county tax roll records, equivalent to 40 samples from each of the selected counties. The mailing of the questionnaire and the timing of implementation was in accordance with the recommendations made by Dillman et al (2014). This included a pre-notice letter, survey questionnaire, follow-up letter, and second-round survey questionnaire. The pre-notice letter preceded the survey questionnaire by five days. The follow-up letter was sent a week after the mailing of the

survey questionnaire. Finally, the second-round survey questionnaire was sent two weeks after the mailing of the follow-up letter. All elements involved in this survey were approved by Auburn University's Institutional Review Board (Document Approval number: 16-184 EX 1605).

Figure 3.1. Counties in Alabama selected to receive the 2016 survey of wild hog damage on non-industrial private lands.



3.2.6. Survey Response Data Analysis and Organization

Task 3 involved performing descriptive statistics on the survey data using R Statistical Analysis Software (R Core Team 2013). Alabama consists of physiographic regions with a variety of properties promoting or limiting the optimal growth of certain tree species. For this reason the response data was organized by major physiographic regions of Alabama which include the Limestone Valley (LV), Appalachian Plateau (AP), Upper Coastal Plains (UCP), Black Belt Prairies (BBP), Piedmont Plateau (PP), and Lower Coastal Plains (LCP). In most instances data were summarized by physiographic region in order to collect averages for questionnaire responses by region. Specific information regarding timber stands planted from 2013-2015 were not summarized by physiographic region in order to obtain State-wide averages for replanting costs due to wild hog damage. Additionally, maps and figures were created to examine the distribution of responses by county using ArcGIS[®] (ESRI 2013).

3.2.7. Overview of Developing a Stand Level Economic Model

For Task 4, the results from the statistical analysis of the survey data were used to develop a stand level economic model utilizing longleaf as the species of interest for the model. Longleaf was selected over loblolly because of its historic association with wild hog damage. The long-term production goal for the stand level economic model was to produce high quality sawtimber and poles on a 60-year rotation (Boyer and Farrar 1981) while utilizing a shelterwood system (Crocker and Boyer 1975). Managing the stand as a shelterwood system would allow natural regeneration to be captured near the end of the rotation and either eliminate or reduce future costs associated with seedling purchases.

An additional short-term production goal for the model was to take advantage of pine straw collection during the peak years of litter fall.

Management decisions for utilizing a shelterwood system were in accordance with the recommendations of Croker and Boyer (1975). According to the authors, prescribed burns should be implemented on a 2-3 year cycle over the duration of the rotation (with the exception of years where regeneration has not become established and is particularly vulnerable to fire). During the 60 year rotation, a preparatory cut would be made at year 40 to reduce the stand's basal area (BA) to 60-70 square feet per acre. This cut would be followed by a seed cut in the 50th year of the rotation to further reduce the stand's BA to 35 square feet per acre. Under this management regime an even-aged longleaf pine forest should successfully produce regeneration for the following rotation as well as produce high quality sawtimber and poles.

3.2.8. Stand Level Economic Model: Cost Assumptions

Site preparation and planting procedures were based on the recommendations from Longleaf Alliance (2007) which outline the steps for successfully establishing planted longleaf. The publication recommended mechanical site preparation prior to planting, followed by a herbaceous release treatment and prescribed burn the following year after planting. The costs associated with site preparation, planting, and stand management (herbaceous release treatment and prescribed burn) were taken from Maggard and Barlow's (2017) averages for Southern forestry practices in 2016. Maggard and Barlow (2017) reported hand planting to be more common than machine planting, so the average labor costs associated with this planting method were incorporated into the

model. Survey participants indicated a preference for planting containerized longleaf seedlings over bareroot; therefore, cost assumptions for containerized longleaf seedlings were used (\$0.20 per seedling to hand plant). The pricing for containerized longleaf seedlings were \$0.20 per seedling based on Rayonier's 2016 seedling prices (Mark Davis, Elberta Nursery Director, Rayonier, personal communication, August 8, 2017). Rayonier was selected as a reputable seedling provider in Alabama from The Longleaf Alliance Grower's Directory (Available at: <http://www.longleafalliance.org/what-we-do/restoration-management/resources/nurseries/grower-listings/Alabama>). Finally, the number of seedlings planted per acre was assumed to be the average planting density reported by survey respondents.

3.2.9. Stand Level Economic Model: Growth and Yield Model Development

Predicting longleaf growth and product yield on a stand level is difficult because a truly satisfactory model has yet to be created (John Kush, personal communication, August 15, 2017). When choosing a growth and yield model it is important to consider the performance of the model, the biological realism, the application environment, and design of the modeling process (Buchman and Shifley 1983). While longleaf growth projection models have been used in scientific literature, no suitable models existed or were available for this project. Consequently, longleaf pine growth and yield values were estimated using a combination of available modeling software, weight tables, and real-world data.

NLongleaf (Matney 1996) software was used to project the growth of three longleaf stands grown for 48, 49, and 50 years. The stands had the same site index (SI)

and trees per acre (TPA) upon initiation. Stands were grown with the assumed SI of 70 feet (base age 50) which would be considered normal for sites within the Coastal Plain. The program was useful to an extent but had age limitations on the thinning operations which could be performed. For this reason the different aged stands were grown to the point where the preparatory cut would take place such that the three stands were 38, 39, and 40 years old. The diameter distribution, TPA, tree heights, and BA were recorded from each stand's projected inventory. The largest and smallest trees were removed from the stand inventory until the remaining BA was between 60-70 square feet per acre. After the preparatory cut was made, useful growth projection information from NLongleaf (Matney 1996) was limited for its application to the projects growth and yield model. NLongleaf does not allow for stand management options after the 37th year of a rotation so growth projections after the preparatory cut could not be generated utilizing this particular software.

Additional data were needed in order to project the growth of the stands an additional 10 and 20 years past the preparatory cut. The data were taken from the Escambia Experimental Forest in south Alabama which was established in 1947. Since that time the longleaf demonstration forest has been managed by periodic thinnings and prescribed fire on a three year cycle. Plots were selected with similar TPA and SI as was assumed for the NLongleaf model using stand table projection methods. Using stand table projection methods trees were then sorted into age groups in order to calculate the growth index ratio (GIR) (Avery and Burkhart 2015). The GIR was used to estimate how the diameter distribution of trees changed over time. The GIR for trees growing from ages 40-50 (n = 43) and 50-60 (n = 11) was the difference in the diameter at breast height

(DBH) for a given tree in each age class after 10 years of growth. The DBH is a tree measurement often used in conjunction with calculating the BA. DBH is defined as the diameter of a tree at 4.5 feet above the ground.

Linear regressions were run using SPSS (IBM Corp. 2015) to create regression equations capable of predicting the 10-year growth rate for trees within the respective age ranges. The resulting regression equations were:

$$Y'_{40-50} = 0.623 + 0.1X_1 \quad (1)$$

$$Y'_{50-60} = 0.085 + 0.092X_2 \quad (2)$$

Y'_{40-50} is the 10-year growth rate of longleaf pine trees growing from age 40 to 50 and X_1 is the DBH of a tree at age 40. Y'_{50-60} is the 10-year growth rate of longleaf pine trees growing from age 50 to 60 and X_2 is the DBH of a tree at age 50. Equation 1 was used to predict the new diameter distribution of trees grown for 10 years after the preparatory cut (after subtracting out 10 percent of assumed tree mortality). The diameter class distribution at age 50 was then used to calculate the BA of the stand.

A shelterwood cut was then made in the 50th year of the rotation to reduce the BA of the stand to 35 square feet per acre. For this cut the largest trees were removed until the desired BA was achieved. Equation 2 was then used to predict the new diameter distribution of trees grown for 10 years after the seed cut (after subtracting out 10 percent of assumed tree mortality). The final harvest of the remaining trees was to take place in year 60 of the rotation.

Calculating the tons of timber removed after the preparatory cut, seed cut, and final harvest was done using a weight table for predicting single tree weight given the diameter class and tree height (Clark and Saucier 1990). Projected tree heights for diameter classes were extracted from the NLongleaf (Matney 2016) outputs run earlier. With the tree heights and diameter classes previously calculated, we then used the weight table to estimate the individual tree weight for each diameter class (Table 2 in Clark and Saucier 1990). The pounds per tree could then be multiplied by the number of trees harvested to attain the tonnage harvested for each diameter class.

Timber products and pricing were based on the 2016 State-wide averages for Alabama (Harrison et al. 2017). Trees with a 6-7 inch DBH were considered pulpwood products and had an average price of \$9.40 per ton. Trees with an 8-11 inch DBH were considered Chip-n-Saw products and had an average price of \$16.81 per ton. Sawtimber and poles would be the expected timber products from trees with a DBH greater than 12 inches and had average prices of \$24.28 and \$49.86 per ton, respectively. It was assumed with proper management, a longleaf stand would be capable of growing a 60/40 pole to sawtimber ratio after 40 years of age (North Carolina Forest Service 2011). This ratio was applied to trees harvested with DBH's greater than 12 inches.

The majority of Alabama landowners with pine forests indicated an interest in taking advantage of non-timber forest products like pine straw (Dyer 2012); therefore, pine straw harvesting was incorporated into the economic model. The model's harvesting schedule was developed in accordance with the recommendations of Franklin (2008) to allow for pine straw to be collected during peak periods of litter fall. Pine straw would be

collected every other year for 12 years starting in year 10 of the rotation. For the purposes of the model, harvesting would not take place on stands younger than 10 years old. In order to calculate pine straw yields a pine straw production function from Dyer (2012) was utilized:

$$Y' = 1.266X1 + -0.661X2 + 1.228X3 + 21.043 \quad (3)$$

where Y' is the number of green bales of pine straw produced per acre per year, X1 is the BA in square feet per acre, X2 is the stand age, and X3 is the SI. The pine straw production equation should be a fairly accurate predictor of pine straw yields for the modeled stands because the growth parameters resembled those to which the equation was intended. Input data used to calculate pine straw yields was based on the projected stand inventory data from Nlongleaf (Matney 1996). The price received per bale of pine straw was assumed to be the average price a landowner would expect to receive according to Franklin (2008) (\$0.35-\$3.00). The price used of \$1.68 per bale was intended to be a fairly modest estimate.

3.2.10. Stand Level Economic Model: Scenarios for Exploring Possible Landowner Responses to Wild Hog Damage

Revenues from timber and pine straw growth and yield models were calculated for stands grown for 58 years (stand-58), 59 years (stand-59), and 60 years (stand-60). Four scenarios were created to examine how profitability of the modeled stand might be affected by wild hog damage given a landowner's response. The net present value (NPV) was calculated for each scenario at a discount rate of 5 percent. In the first scenario

(scenario-1), no damage occurs, thus replanting is not necessary. The NPV for a full rotation or stand-60 is used to calculate the economic value of the forest stand for scenario-1. Survey results indicated average seedling mortality of 23 percent for acres impacted by wild hogs; therefore, this percent of damage was used for scenarios 2-4. In the second scenario (scenario-2), the modeled stand was damaged by wild hogs and the landowner has decided to replant the damaged portion the following year. The seedlings replanted a year later have one less year of growth than the seedlings planted at the beginning of the rotation. In order to calculate how this 1-year loss of growth would affect timber and pine straw revenues, the following equation was used:

$$\text{Revenues}_{\text{scenario2}} = ((1 - \% \text{damage}) \text{Revenue}_{\text{stand-60}}) + (\% \text{damage} \times \text{Revenue}_{\text{stand-59}}) \quad (4)$$

where $\text{Revenues}_{\text{scenario2}}$ is the discounted timber and pine straw revenue for scenario-2, $\% \text{damage}$ is the seedling mortality from wild hog predation, $\text{Revenue}_{\text{stand-60}}$ is the discounted timber and pine straw revenues for stand-60, and $\text{Revenue}_{\text{stand-59}}$ is the discounted timber and pine straw revenues for stand-59. Replanting costs based on the percent of the stand damaged were also incorporated when replanting took place. The difference between the NPV's from scenario-1 and -2 represent the cost of wild hog damage to the stand that could not be recovered even after replanting the next year. It could alternatively be viewed as the amount a landowner could invest in wild hog control given an expected amount of damage they thought might occur.

In scenario-3, we examined how the forest stand would be affected if wild hog damage occurred but replanting was delayed for two years. In order to calculate how a two year loss of growth would affect timber and pine straw revenues, equation 4 was

used but the revenues from stand-58 were substituted for stand-59. The difference between the NPV's from scenario-1 and -3 represent the cost of wild hog damage which could not be recovered with replanting delayed an extra year. It could also alternatively be viewed as the amount a landowner could invest in wild hog control given an expected amount of damage they thought might occur.

In the final scenario (scenario-4), we examined the cost of damage to a forest stand where wild hog damage occurred but replanting did not take place. This was calculated by subtracting out the amount of seedling mortality from the timber and pine straw revenues from stand-60. Because replanting did not occur, the costs associated with scenario-1 would be the same as those for scenario-4. The difference between the NPV's for scenario-1 and -4 represent the loss in value of the forest stand caused by wild hog damage.

3.2.11. Conducting Sensitivity Analyses on Scenarios 2-4

In Task 5, sensitivity analyses were performed on scenarios 2-4 to examine how the cost of wild hog damage fluctuated when exposed to a range of discount rates and amounts of damage. The cost of damage can also be interpreted as the amount which could be alternately invested in control methods. In order to give landowners an idea of what it would cost to reduce wild hog damage in their young forest plantation during periods when seedlings are in a vulnerable stage of growth, three levels of control were created based on different combinations of the most commonly used control methods reported from survey respondents. For the lowest level of control only shooting on site and hunting (without dogs) were combined. A medium level of control combined

shooting on site, hunting (without dogs), and trapping. A high level of control utilized shooting on site, hunting (without dogs), trapping, and hunting with dogs. The annual spending on each of the control methods was based on the average spending reported by survey respondents. The required years of implementing control methods varied depending on if the landowner replanted. The cost of control was calculated for 3, 4, and 5 years of implementation at each level of intensity. Based on the historic reports, wild hog predation seems to primarily occur among very young pine seedlings in or just leaving the grass stage. It was assumed after three growing seasons the planted longleaf seedlings would have left the grass stage and developed enough so they were less appealing to wild hogs. Based on when replanting occurred, vulnerable seedlings could remain in the plantation up to fifth year of the rotation.

3.3. Results

3.3.1. Survey Results

Out of the 1,200 surveys mailed, 406 landowners responded which resulted in 352 usable surveys. The response rate for this survey was 35 percent after deducting 40 surveys from the total mailed out due to bad addresses, deceased recipients, or those declining to participate. A summary of responses by county based on physiographic region is listed in Table 3.1 and illustrated in Figure 3.2. It should be noted that a number of responses came from 15 additional counties not in the original sample group but were included in the analysis because the data still proved pertinent. These respondents owned properties outside of the county with their listed mailing address. The counties with the highest response rates ($n = 17$) were Bullock (Black Belt Prairie), Coffee (Lower Coastal Plain), Sumter (Black Belt Prairie), and Tallapoosa (Piedmont Plateau) Counties. The Lower Coastal Plains had the most counties represented by survey responses ($n = 11$) compared to other regions.

Table 3.1. Counties and the number of corresponding responses (in parentheses) to the 2016 survey of wild hog damage to non-industrial private lands in Alabama. Counties are organized by major physiographic region.

Limestone Valley	Appalachian Plateau	Upper Coastal Plain	
Cherokee (1)	Blount (8)	Autauga (12)	Tuscaloosa (9)
Colbert (10)	Etowah (10)	Bibb (9)	
Lawrence (6)	Jackson (2)	Chilton (1)	
Limestone (11)	Jefferson (1)	Fayette (6)	
Madison (13)	Marshall (9)	Lamar (9)	
	Morgan (7)	Marion (12)	
	Shelby (8)	Pickens (1)	
Black Belt Prairie	Piedmont Plateau	Lower Coastal Plain	
Bullock (17)	Cleburne (1)	Butler (2)	Geneva (2)
Dallas (15)	Clay (1)	Choctaw (14)	Henry (3)
Lowndes (3)	Coosa (2)	Clarke (10)	Houston (14)
Macon (1)	Lee (14)	Coffee (17)	Monroe (2)
Marengo (1)	Tallapoosa (17)	Conecuh (1)	Washington (9)
Montgomery (10)		Covington (9)	Wilcox (16)
Sumter (17)		Crenshaw (9)	

Figure 3.2. Distribution and number of responses to the 2016 survey of wild hog damage to non-industrial private lands in Alabama. Counties are organized by major physiographic region.

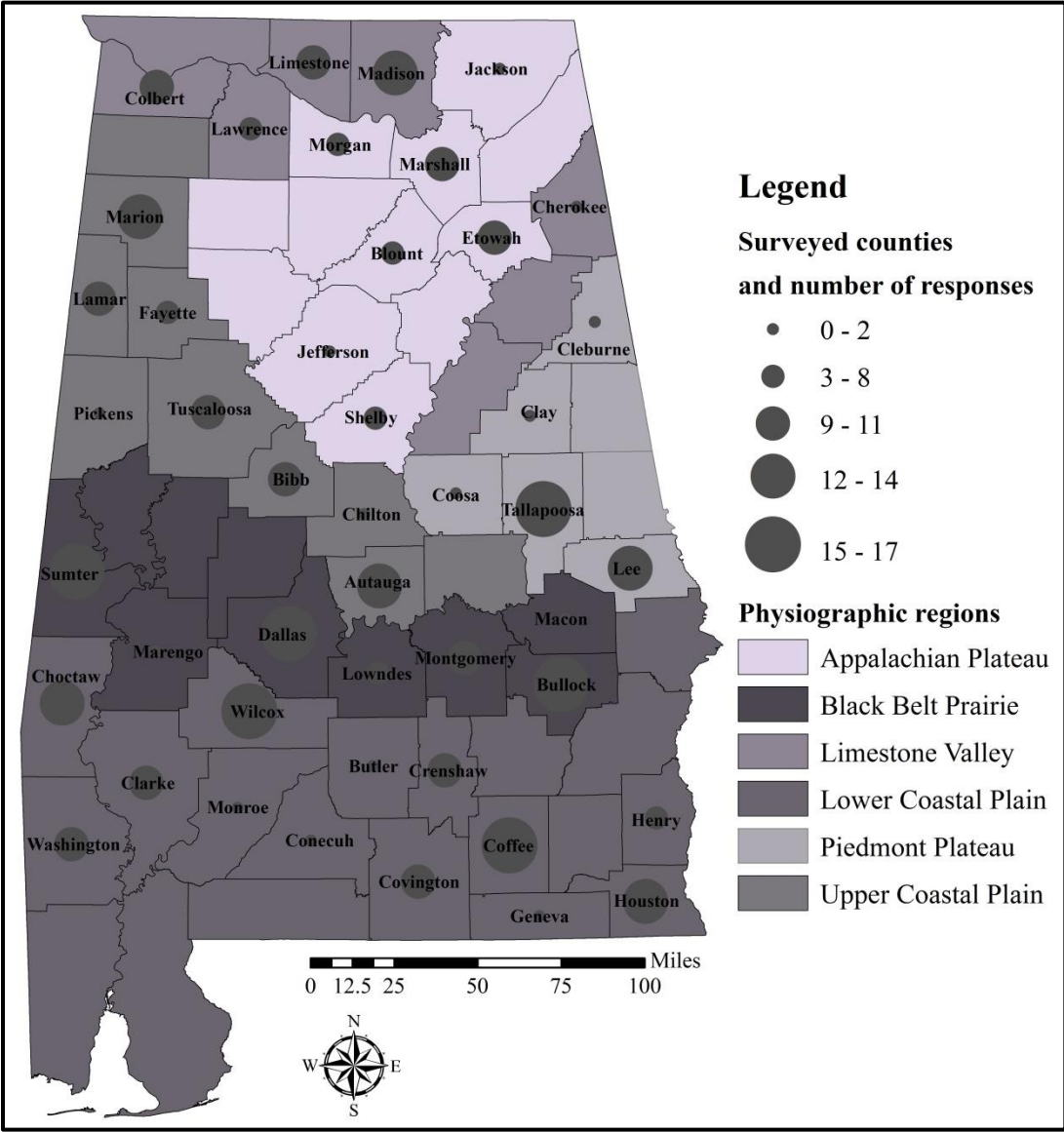


Table 3.2 summarizes the number of responses by region and other relative information about forest plantations, damage, and control efforts. Responses were evenly distributed among regions except for the LCP which had nearly double the response rate compared to other regions. The BBP and LCP held the highest reported values for all categories listed in the table. The largest sample of young forest plantations initiated in 2013-2015 were reported from the LCP ($n = 25$). The number of respondents who planted seedlings during 2013-2015 made up 17 percent of the total respondents who returned usable surveys. The percentage of respondents reporting wild hog damage to their stand was similar between the BBP (33.3%) and the LCP (36.0%). Of the young forest plantations planted from 2013-2015, 25 percent were reported to have sustained damage from wild hogs. Survey results indicate the BBP and the LCP had the largest sample of landowners reporting wild hogs on their land and use of control methods. The regions with the smallest sample of landowners reporting wild hogs on their land and use of control methods were the LV and AP. The highest numbers of hogs killed in 2015 were reported from the BBP, LCP, and LV; in fact, there were a number of cases where respondents indicated killing over 200 wild hogs on their property. A number of respondents indicated they had killed wild hogs on their property but did not report the quantity; therefore, averages in the last column of the table only consider respondents who reported a positive number of wild hogs killed.

Table 3.2. Results from the 2016 survey of non-industrial private landowners. Total responses and percent for each region of Alabama reporting a forest plantation, damage from wild hogs, and control.

Physiographic region	Responses (#)	Landowners who planted seedlings between 2013-15 (%)	Forest stand damage by wild hogs (%)	Wild hogs on land (%)	Attempt to control wild hogs (%)	Hunt on property (%)	Wild hogs killed in 2015 (of positive) (#)
Limestone Valley	41	4	25.0%	7.7% (39)	6.1% (33)	69.7% (33)	125 (1)
Appalachian Plateau	45	2	0.0%	7.5% (40)	3.1% (32)	47.1% (34)	3.3 (2)
Upper Coastal Plain	59	9	0.0%	26.3% (57)	10.0% (50)	79.6% (49)	5.8 (4)
Black Belt Prairie	64	15	33.3%	70.5% (61)	49.1% (57)	81.8% (55)	72 (24)
Piedmont Plateau	35	5	0.0%	17.1% (35)	18.2% (33)	83.9% (31)	1.3 (3)
Lower Coastal Plain	108	25	36.0%	53.8% (106)	36.5% (96)	81.4% (97)	22.8 (39)

Numbers of responses in parentheses where necessary

The results of seedling mortality due to wild hog activity and the associated costs with replanting are presented in Table 3.3. The top three tree species planted by landowners across the State from 2013-2015 were loblolly (n = 46), longleaf (n = 35), and unspecified oak (n = 9). For pine stands planted in 2013-2015, the average number of acres planted was 130 and 47 for longleaf and loblolly, respectively. The greatest amount of forest damage from wild hogs was reported to have occurred in longleaf plantations. Where damage occurred, wild hogs destroyed more loblolly seedlings (179 seedlings/acre) than longleaf (125 seedlings/acre) but affected larger portions of longleaf plantations (34 percent) than loblolly (13 percent). Respondent's answers did not specify how wild hogs had destroyed the seedlings so it could not be determined if seedling mortality was due to direct predation. It was reported the average longleaf plantation had 44 acres damaged by wild hogs. On damaged acres, wild hogs destroyed an average of 23 percent of longleaf seedlings. Loblolly plantations had an average of 6 acres damaged; for every acre damaged in loblolly plantations, wild hogs destroyed an average of 27 percent of seedlings. Assuming replanting occurred, we calculated the cost to replant areas affected by wild hogs in longleaf and loblolly stands. On a per acre basis, replanting costs for damaged longleaf acres were nearly double that for loblolly due to the higher seedling and planting costs associated with containerized longleaf seedlings. The total costs to recover damaged portions of the average sized stands were calculated to be \$2,185.92 and \$150.14 for longleaf and loblolly, respectively. No wild hog damage was reported in any of the hardwood plantations.

Table 3.3. Results from the timber stand section of the 2016 survey of non-industrial private landowners in Alabama. Reported below are the averages for the forest stand information, wild hog damage, and calculated replanting cost assuming respondents replanted after damage occurred.

Species	Number reporting	Size of stand (Acres)	Trees per acre planted	Acres damage by wild hogs (Acres)	Percent of seedlings destroyed per damaged acre (%)	Replanting cost per acre (\$/acre damaged)	Replanting cost for the average stand (\$)
Longleaf pine	32	130	540	44	23	*49.68	*2,185.92
Loblolly pine	41	47	662	6	27	*25.02	*150.14
Oak	8	15	344	0	-	-	-

*Planting cost are based on Maggard and Barlow (2017), seedling cost are based on Rayonier's 2016 prices in Alabama: \$0.40 and \$0.14 per seedling for longleaf and loblolly, respectively.

The top three land cover types with the most frequent sightings of wild hogs were agricultural fields, grassland/pastureland, and bottomland hardwoods (Appendix A). Survey results indicate that of the land cover options, hardwood plantations had the least number of wild hog sightings.

The total cost of each control method by region is presented in Table 3.4. Regardless of the region, the top four most common control methods reported were hunting (without dogs), shooting on sight, trapping, and hunting with dogs. Only two respondents in the BBP region indicated the use of repellents for wild hog control. Based on the reports of responding landowners, the use of fencing (electric or non-electric) for controlling wild hog damage was relatively uncommon. Those who used electric or non-electric fencing made up less than 3 percent of landowners. The largest sample of landowners who reported the use of electric (n=5) and non-electric fencing (n=4) for wild hog control were in the LCP and BBP, respectively. Of the regions considered, landowners in the BBP spent the highest total amount on shooting on sight (\$6,450), hunting without dogs (\$6,370), and trapping (\$20,500). The highest regional total spent on hunting with dogs was \$2,550 by landowners in the LCP. No monetary values were reported for wild hog control from the AP.

Table 3.4. Total cost of wild hog control methods by physiographic region in Alabama from the 2016 survey of non-industrial landowners.

Physiographic region	Shoot on sight	Hunt w/dogs	Hunt w/out dogs	Aerial	Trap	Repellents	Electric fence	Non-electric fence
Limestone Valley	\$60 (2)	\$0 (0)	\$0 (2)	\$0 (0)	\$2,300 (2)	\$0 (0)	\$0 (1)	\$0 (0)
Appalachian Plateau	\$0 (1)	\$0 (0)	\$0 (0)	\$0 (0)	\$0 (0)	\$0 (0)	\$0 (0)	\$0 (0)
Upper Coastal Plain	\$500 (4)	\$0 (0)	\$0 (1)	\$0 (0)	\$1,200 (3)	\$0 (0)	\$0 (0)	\$0 (1)
Black Belt Prairie	\$6,450 (25)	\$0 (9)	\$6,370 (16)	\$0 (0)	\$20,500 (23)	\$5,500 (2)	\$233 (1)	\$1,233 (4)
Piedmont Plateau	\$30 (5)	\$0 (2)	\$0 (4)	\$0 (0)	\$0 (1)	\$0 (0)	\$0 (0)	\$0 (0)
Lower Coastal Plain	\$3,070 (36)	\$2,550 (12)	\$250 (21)	\$0 (0)	\$9,915 (18)	\$0 (0)	\$1,387 (5)	\$0 (0)

Numbers of responses in parentheses

Fencing costs are annualized based on initial cost and suspected length of use

The average amount spent annually on control methods in each region is listed in Table 3.5. With the highest cost of those reported for regions offering larger sample sizes, a landowner in the BBP spent, on average, \$258 shooting on sight, \$398 hunting without dogs, and \$891 trapping. The LCP had the largest sample size and highest average expense paid by landowners to hunt with dogs at \$212. The percentage of landowners in each region who indicated using control methods is reported in Table 3.6. The only regions which utilized hunting with dogs were the BBP, LCP, and PP. The BBP was the only region with reports of all control methods being utilized by landowners with the exception of aerial hunting. For this region the most commonly implemented methods were shooting on sight (43.9%) and trapping (40.4%).

Table 3.5. Average spending of wild hog control methods by physiographic region in Alabama from the 2016 survey of non-industrial landowners.

Physiographic region	Shoot on sight	Hunt w/dogs	Hunt w/out dogs	Aerial	Trap	Repellents	Electric fence	Non-electric fence
Limestone Valley	\$30 (2)	-	\$0 (2)	-	\$1,150 (2)	-	\$0 (1)	-
Appalachian Plateau	\$0 (1)	-	-	-	-	-	-	-
Upper Coastal Plain	\$125 (4)	-	\$0 (1)	-	\$400 (3)	-	-	\$0 (1)
Black Belt Prairie	\$258 (25)	\$0 (9)	\$398 (16)	-	\$891 (23)	\$2,750 (2)	\$233 (1)	\$308 (4)
Piedmont Plateau	\$6 (5)	\$0 (2)	\$0 (4)	-	\$0 (1)	-	-	-
Lower Coastal Plain	\$85 (36)	\$212 (12)	\$12 (21)	-	\$551 (18)	-	\$277 (5)	-

Numbers of responses in parentheses

Fencing costs are annualized based on initial cost and suspected length of use

Table 3.6. Percentage of responses reporting use of wild hog control methods by physiographic region in Alabama from the 2016 survey of non-industrial landowners.

Physiographic region	Shoot on sight	Hunt w/dogs	Hunt w/out dogs	Aerial	Trap	Repellents	Electric fence	Non-electric fence
Limestone Valley	6.1%	0.0%	6.1%	0.0%	6.1%	0.0%	16.7%	0.0%
Appalachian Plateau	3.1%	0.0%	0.0%	0.0%	0.0%	0.0%	0.0%	0.0%
Upper Coastal Plain	8.0%	0.0%	2.0%	0.0%	6.0%	0.0%	0.0%	16.7%
Black Belt Prairie	43.9%	15.8%	28.1%	0.0%	40.4%	3.5%	16.7%	66.7%
Piedmont Plateau	15.2%	6.1%	12.1%	0.0%	3.0%	0.0%	0.0%	0.0%
Lower Coastal Plain	37.5%	12.5%	21.9%	0.0%	18.8%	0.0%	83.3%	0.0%

The majority of respondents answering the WTP/WTa question indicated a preference for wild hog eradication (Table 3.7). Landowners preferring wild hogs on their property indicated a dollar amount they would have to be paid to accept eradication (WTa), while those in favor of permanently getting rid of wild hogs indicated an amount they are willing to pay for eradication (WTP). From this sample, landowners who wanted fewer/no wild hogs had an willingness to pay for eradication that exceeded the amount required by those who prefer wild hogs but were willing to be financially supplemented to allow their eradication, such that $WTP > WTa$.

Only one respondent indicated wanting to be supplemented for accepting eradication. Of those who preferred wild hogs on their property, 56 percent hunted hogs but none reported generating income from hog hunting leases. Only two individuals in favor of eradication reported any profitability from hog hunting. The income generated from their hog hunting leases made up 2.2 percent of the total amount made from all hunting leases.

Of the landowners who indicated a dollar amount they would be willing to pay for eradication, 43 percent said they would be willing to pay while others wanted fewer/no wild hogs but were not willing to pay for wild hog eradication. The survey data suggests a trend that landowners without wild hogs on their property were willing to pay more than those with wild hogs. On a regional basis, landowners in the BBP who did not have wild hogs on their land were willing to pay the most (\$206 per acre per year) for eradication (Table 3.8).

Table 3.7. Alabama landowners' mean reported willingness to pay (WTP) for or willingness to accept (WTA) wild hog eradication (\$ per acre/year) from the 2016 survey of non-industrial private landowners.

	Pigs on land	n	Mean (\$)
WTP for eradication	N	151	\$22.40
(prefer fewer/none)	Y	114	\$8.37
WTA eradication	N	5	-
(prefer them)	Y	9	\$71.43

Table 3.8. Mean reported willingness to pay (WTP) for wild hog eradication (\$ per acre/year) by physiographic region in Alabama from the 2016 survey of non-industrial private landowners.

Physiographic region	Pigs on land	n	Mean (\$)
	N	2	\$1.09
Limestone Valley	Y	23	\$5.00
	N	24	\$14.38
Appalachian Plateau	Y	2	\$5.00
	N	22	\$3.86
Upper Coastal Plain	Y	16	\$9.69
	N	10	\$206.00
Black Belt Prairie	Y	29	\$11.03
	N	15	\$3.33
Piedmont Plateau	Y	2	\$2.50
	N	26	\$5.58
Lower Coastal Plain	Y	47	\$6.81

3.3.2. Stand Level Economic Model: Growth and Yield Results for Projected Timber and Pine Straw Revenues

The timber harvest results from the modeled stands after each cut are presented in Table 3.9. Stand-58 (forest stand with 58 year rotation) had the highest weight of timber removed throughout the rotation yet had the lowest net present value (NPV) for timber revenue. This can be explained by stand-58 consisting of a higher proportion of smaller trees compared to the other stands. Stand-59 (forest stand with 59 year rotation) had the second highest weight of timber removed and a NPV for timber revenue \$1.10 more than the NPV for stand-60's (forest stand with 60 year rotation) timber harvest. Stand-60 held a higher proportion of larger trees compared to other stands which resulted in the greatest quantity of sawtimber and pole products being produced. A two years loss in rotation length resulted in an 11 percent decrease in sawtimber and pole products for stand-58 compared to stand-60. The greatest difference in NPV's for timber revenues was seen between stand-60 and stand-58, which was approximately \$12 per acre.

The pine straw harvest results for the modeled stands are presented in Table 3.10. Stand-60 produced 135 and 157 more bales of pine straw per acre per year than stand-59 and stand-58, respectively. The year loss in rotation length between stand-60 and stand-59 resulted in a difference in NPV for pine straw sales of \$132 an acre. A two year loss in rotation length between stand-60 and stand-58 resulted in a \$149 an acre loss of potential pine straw revenue for stand-58.

Table 3.9. Stand level economic model: projected timber harvest for 3 stands of varying age utilizing a shelterwood system growing longleaf pine on a 60-year rotation given a site index of 70 and planted at 540 trees per acre.

	Pulp (tons/acre)	C-N-S (tons/acre)	Sawtimber (tons/acre)	Poles (tons/acre)	Total (tons/acre)	NPV (\$)
Preparatory cut, year 40						
Stand-60	14.16	0	4.32	6.48	24.96	\$ 79.69
Stand-59	14.69	0	3.82	5.73	24.24	\$ 73.38
Stand-58	15.48	0	2.73	4.10	22.31	\$ 59.10
Seed cut, year 50						
Stand-60	0	0	15.80	23.71	39.51	\$ 136.53
Stand-59	0	2.86	16.60	24.89	44.35	\$ 147.56
Stand-58	0	5.51	16.55	24.83	46.89	\$ 151.06
Final harvest, year 60						
Stand-60	0	24.76	6.34	9.51	40.61	\$ 55.93
Stand-59	0	26.59	5.35	8.03	39.97	\$ 52.30
Stand-58	0	28.62	4.56	6.84	40.02	\$ 49.93

Product	DBH range	\$/ton
Pulpwood (Pulp)	6-7 inch	\$ 9.40
Chip-n-saw (C-N-S)	8-11 inch	\$ 16.81
Sawtimber	12+ inch	\$ 24.28
Poles	12+ inch	\$ 49.86

Table 3.10. Stand level economic model: projected pine straw harvest for 3 stands of varying age with harvest occurring every other year for 12 years starting at year 10 of the rotation given a site index of 70 and planted at 540 trees per acre.

	Total green bales harvested	\$/bale	NPV
Stand-60	937	\$ 1.68	\$ 717.30
Stand-59	802	\$ 1.68	\$ 585.12
Stand-58	780	\$ 1.68	\$ 568.49

3.3.3. Stand Level Economic Model: Economic Results of Four Scenarios for an Even Aged Longleaf Pine Operation Utilizing a Shelterwood System

Table 3.11 presents the results of the four scenarios created to explore the potential economic impact of wild hogs to a longleaf pine operation. A sub-table is also included with the cost assumptions used to calculate the NPV for each scenario. Survey results indicated that acres damaged by wild hogs had 23 percent seedling mortality; therefore, this percent was applied to scenarios where wild hog damage occurred. Additionally, a five percent discount rate was also assumed for each of scenario's NPV calculations.

The NPV for scenario-1 was \$430 per acre with \$560 per acre in establishment and maintenance costs and \$989 per acre in discounted timber and pine straw revenues. In scenario-2, pine straw and timber revenues decreased by \$30 per acre while replanting added \$47 per acre more to the operation's costs resulting in a NPV of \$352 per acre. The cost of the damage that could not be recovered by replanting the following year was \$77 per acre. Alternatively, if a landowner was considering establishing a longleaf plantation in an area where they expected wild hogs to damage at least 23 percent of seedlings per acre then they could allocate \$77 per acre to be spent on control methods.

In scenario-3, pine straw and timber revenues decreased an additional \$7 an acre with a resulting NPV of \$348 per acre. The cost of the damage from delaying replanting an additional year was \$82 per acre. In the final scenario, establishment and maintenance costs did not change from scenario-1 because replanting did not occur; however, the timber and pine straw revenues decreased \$227 per acre as a result of the wild hog

damage bringing the NPV to \$202 an acre. The resulting cost of damage was the greatest for scenario-4 at \$228 per acre. Based on these results, if a landowner replanted the year after damage occurred they could recover nearly 66 percent of the value that would be lost if they did not replant. Alternatively, if a landowner decided to delay replanting by an additional year they could still recover 64 percent of the value that would be lost if they did not replant at all.

Table 3.11. Stand level economic model: the costs, revenues, and net present values for four scenarios involving wild hog's potential impact on an even-aged planted longleaf pine operation on a 60 year rotation utilizing a shelterwood system.

	Revenues (\$/ac)	Costs (\$/ac)	Damage (%/ac)	NPV (\$/ac)	Cost of damage (\$/ac)
Scenario 1: no damage	\$ 989	\$ 560	0%	\$ 430	\$ -
Scenario 2: damage, replant	\$ 959	\$ 607	23%	\$ 352	\$ 77
Scenario 3: damage, delay replant	\$ 952	\$ 605	23%	\$ 348	\$ 82
Scenario 4: damage, no replant	\$ 762	\$ 560	23%	\$ 202	\$ 228

Assumptions:	
Costs	
Site preparation	
Mechanical	\$141/ac
Herbicide	\$48/ac
Prescribed burn	\$155/ac
Planting	
Seedlings	\$108/ac
Hand planting	\$108/ac
Discount rate	5%

3.3.4. *Sensitivity Analyses Results*

Sensitivity analyses were performed on scenarios 2-4 to examine how the cost of wild hog damage fluctuated when exposed to a range of discount rates and amounts of damage. The discount rates ranged from 2-6 percent to cover all realistic discount rates a landowner might expect for a forestry investment. Amounts of wild hog damage ranged from 3-48 percent increasing by increments of five percent. The results of the analyses are presented in Tables 3.12-3.14. The sensitivity analyses revealed slight adjustments in discount rates and amounts of damage sustained impacted the potential profitability of the forest stand.

The resulting cost of damage caused by wild hogs from scenarios 2-4 represent the value a landowner could alternately spend on reducing future damage. Without a frame of reference on what the control costs might be these values mean little. In order to estimate how much control could be afforded given the cost of damage, sub-tables were included below the sensitivity analyses tables. The levels of control in the sub-tables were calculated over a range of discount rates corresponding with the sensitivity analyses for the cost of damage. For example, if a landowner predicts a 23 percent loss of seedlings per acre due to wild hogs, but plans to replant the next year, they could expect the damage to cost them \$77 an acre (at a 5 percent discount rate) (Table 3.12). The \$77 per acre could have been spent controlling wild hogs to prevent or reduce future damage. If replanting occurred after the first year of the rotation, seedlings vulnerable to hog predation would remain in the plantation until after the fourth year. Given this scenario, a landowner could only afford the lowest level of control spending (shooting on site and

hunting without dogs). If the landowner did not replant after damage occurred but wanted to protect the remaining seedlings, they could invest the expected value for the cost of damage in the highest level of control.

The results of the sensitivity analyses suggest a landowner considering establishing a longleaf plantation would benefit from investing the equivalent to a high amount of wild hog control to prevent or reduce damage during the first three growing seasons. If damage does occur, replanting the following year is the viable option to recover the majority of value lost to wild hog damage. By delaying replanting an additional year a landowner only incurs an additional \$5 to the cost of damage from wild hogs.

Table 3.12. Scenario 2: Sensitivity analysis- estimated cost of wild hog damage (\$/ac) to a young longleaf pine plantation given that replanting takes place the year after damage occurs.

		Discount rate				
%damage,		2%	3%	4%	5%	6%
3%	\$	12	\$ 11	\$ 11	\$ 10	\$ 10
8%	\$	31	\$ 30	\$ 28	\$ 27	\$ 26
13%	\$	50	\$ 48	\$ 46	\$ 44	\$ 42
18%	\$	70	\$ 66	\$ 63	\$ 61	\$ 58
23%	\$	89	\$ 85	\$ 81	\$ 77	\$ 74
28%	\$	108	\$ 103	\$ 99	\$ 94	\$ 90
33%	\$	128	\$ 122	\$ 116	\$ 111	\$ 106
38%	\$	147	\$ 140	\$ 134	\$ 128	\$ 123
43%	\$	166	\$ 159	\$ 151	\$ 145	\$ 139
48%	\$	186	\$ 177	\$ 169	\$ 162	\$ 155

Sub-Table 3.12. Cost (\$/ac) for 4 years of varying levels of wild hog control exposed to a range of discount rates.

		Discount rate				
Control level,		2%	3%	4%	5%	6%
High	\$	118	\$ 115	\$ 113	\$ 110	\$ 107
Medium	\$	99	\$ 97	\$ 95	\$ 93	\$ 90
Low	\$	56	\$ 55	\$ 54	\$ 52	\$ 51

Table 3.13. Scenario 3: Sensitivity analysis- estimated cost of wild hog damage (\$/ac) to a young longleaf pine plantation given that replanting is delayed until 2 years after damage occurs.

Discount rate						
%damage,	2%	3%	4%	5%	6%	
3%	\$ 14	\$ 12	\$ 12	\$ 11	\$ 10	
8%	\$ 36	\$ 33	\$ 31	\$ 29	\$ 27	
13%	\$ 59	\$ 54	\$ 50	\$ 46	\$ 43	
18%	\$ 82	\$ 75	\$ 69	\$ 64	\$ 60	
23%	\$ 105	\$ 96	\$ 88	\$ 82	\$ 77	
28%	\$ 127	\$ 116	\$ 107	\$ 100	\$ 94	
33%	\$ 150	\$ 137	\$ 127	\$ 118	\$ 110	
38%	\$ 173	\$ 158	\$ 146	\$ 136	\$ 127	
43%	\$ 195	\$ 179	\$ 165	\$ 153	\$ 144	
48%	\$ 218	\$ 199	\$ 184	\$ 171	\$ 160	

Sub-Table 3.13. Cost (\$/ac) for 5 years of varying levels of wild hog control exposed to a range of discount rates.

Discount rate						
Control level,	2%	3%	4%	5%	6%	
High	\$ 146	\$ 142	\$ 138	\$ 134	\$ 131	
Medium	\$ 123	\$ 120	\$ 116	\$ 113	\$ 110	
Low	\$ 70	\$ 68	\$ 66	\$ 64	\$ 62	

Table 3.14. Scenario 4: Sensitivity analysis- estimated cost of wild hog damage (\$/ac) to a young longleaf pine plantation given that no replanting occurs after damage.

		Discount rate				
%damage,		2%	3%	4%	5%	6%
3%	\$	69	\$ 50	\$ 38	\$ 30	\$ 24
8%	\$	183	\$ 134	\$ 102	\$ 79	\$ 63
13%	\$	298	\$ 218	\$ 165	\$ 129	\$ 103
18%	\$	413	\$ 302	\$ 229	\$ 178	\$ 142
23%	\$	527	\$ 386	\$ 292	\$ 228	\$ 182
28%	\$	642	\$ 470	\$ 356	\$ 277	\$ 221
33%	\$	757	\$ 554	\$ 419	\$ 327	\$ 261
38%	\$	871	\$ 638	\$ 483	\$ 376	\$ 300
43%	\$	986	\$ 722	\$ 546	\$ 425	\$ 340
48%	\$	1,101	\$ 806	\$ 610	\$ 475	\$ 379

Sub-Table 3.14. Cost (\$/ac) for 3 years of varying levels of wild hog control exposed to a range of discount rates.

		Discount rate				
Control level,		2%	3%	4%	5%	6%
High	\$	89	\$ 88	\$ 86	\$ 84	\$ 83
Medium	\$	75	\$ 74	\$ 72	\$ 71	\$ 70
Low	\$	43	\$ 42	\$ 41	\$ 40	\$ 39

3.4. Discussion

The primary goal of this project was to estimate the cost of wild hog damage to non-industrial private landowner's forest stands in Alabama. In order to accomplish this goal we used a questionnaire to solicit pertinent information needed to create a damage estimate. Survey results indicated wild hog's damaged more acres in longleaf plantations than any other tree species planted from 2013-2015. The resulting cost to replant damaged acres of the average longleaf plantation was 15 times more than replanting cost for the average loblolly plantation (based on Rayonier's 2016 seedling prices and planting costs from Maggard and Barlow 2016). The higher replanting costs associated with longleaf are primarily due to the increased cost of seedlings. Containerized longleaf seedlings cost 65% more than bareroot loblolly seedlings. Due to the higher seedling cost and wild hog damage associated with longleaf, we recommend landowners invest in wild hog control to protect their plantation during the first three growing seasons. The estimated cost of wild hog damage extends beyond the cost of replanting and includes the amount that must be invested in preventing future damage (i.e. wild hog control), for this reason additional economic and sensitivity analysis was necessary.

In addition to estimating the cost of wild hog damage to forestry operations on non-industrial private lands, this project also included the economic analysis of potential management decisions in response to damage. If a landowner establishes a longleaf plantation in an area where wild hogs are known to reside, then they should be aware of the potential risk of damage occurring. While being proactive with wild hog control does not guarantee damage will not occur, it may prove a viable option for mitigating the risk

of damage compared to a more reactive approach. Investing in wild hog control upfront could prevent the time, effort, and additional costs associated with having to replant. Even so, scenario-2 showed that more immediate replanting is the best way to reduce the cost of the damage to the landowner. Scenario-3 may be the good option in some cases where a landowner would want to invest additional funds and time into wild hog control while still recovering a portion of the value lost to damage.

Due to the long-term nature of forestry investments, landowners are likely to want to replant even after relatively low level damage occurs. Crop failure in a young forest plantation can offset a rotation by years and carries considerably more economic loss into the future compared to agricultural crops. The cost of damage to a forest stand grows quickly when replanting is not an option as was seen in the sensitivity analysis for scenario-4 (Table 3.14.). For example, with a 5 percent discount rate the cost of 8 percent damage to a longleaf pine plantation would be \$84 per acre (Table 3.14.) when replanting does not take place compared to \$27 per acre (Table 3.12.) with replanting. The decision to replant, invest in control methods, or do nothing will ultimately depend on the landowner and their objectives, but hopefully the tables created for this project will aid in the decision making process.

Survey respondents reported nearly double the amount of acres planted with longleaf than loblolly between 2013 and 2015 in Alabama. Cost-share programs which incentivize landowners to plant longleaf could explain why so many more acres were planted with longleaf than loblolly. Maggard and Barlow's (2017) survey of forestry practices in the South, which is conducted every two years, also reported an increase in a

number of containerized longleaf being planted. If this trend continues it is likely wild hog damage to longleaf plantations will become more common in the South.

Our sample group of respondents consisted of 352 private landowners, of which 17 percent reported establishing a forest stand in 2013-2015. For landowners who planted trees during this time, 25 percent had wild hog damage to their young forest stand. All reports of wild hog damage to forest stands came from counties in the BBP, LCP, or LV. Forestry insurance policies do not cover damages from wild hogs. Landowners who received damage were likely to have had to cover the cost to replant; consequently, forest plantation owners should budget for replanting when wild hogs are present. The NRCS cost-share program provides financial assistance to help cover the costs of site preparation, seedlings costs, and planting of longleaf. The program will help cover the cost of replanting only if the cause of damage was from an “act of nature”. There was at least one case in Alabama in 2016 where the NRCS provided financial assistance to replant after wild hog damage occurred (Tim Albritton, State Forester, NRCS, personal communication, August 25, 2017). In this case, wild hogs were considered an act of nature and the landowner had demonstrated using wild hog control efforts to try to prevent damage. If wild hog populations continue to grow and expand unimpeded, perhaps future conditions will warrant forestry insurance or cost-share program policies to cover wild hog damage.

It is interesting to note landowners reported spending the most on trapping as a method for controlling wild hogs yet assigned it the lowest efficiency rating. It is possible landowners did not feel they were getting a high enough return on their

investment in terms of the hogs killed per dollar spent on trapping. The results also suggested landowners felt hunting with dogs was the most efficient control method. Campbell and Long (2009) conducted an extensive literature review of wild hog damage and damage management in forested ecosystem and surmised that wild hog management is most successful when a variety of techniques are used in an integrated fashion. Wild hogs are very adaptable and can learn to evade capture or removal when only one technique is utilized (Choquenot et al. 1999); therefore, despite how landowners felt about the efficiency of certain control techniques, using a combination of methods should yield the best results.

Creating policies governing invasive species management and use is often a reflection of the costs and benefits individuals associate with their presence. These policies in turn, drive differences in legal classifications, dictate permissible control methods, conservation outcomes, and wildlife densities across the state. The results of this current study can be used to inform policy makers; for example, in deciding which counties would be eligible to receive NRCS assistance, distribution data would be critical to know where the largest populations of wild hogs resided. Data from studies like Conely et al. (2014), who conducted a survey to map the distribution of wild hogs in Alabama, would serve as very beneficial information in deciding where to allocate funding. The group found wild hogs were present in 64 counties and abundant in 31 counties. Our survey also indicated that wild hog populations were distributed through all major physiographic regions in Alabama but appeared to have the greatest densities in the BBP, LCP, and UPC. Interestingly, all 16 counties eligible to receive NRCS assistance funding for wild hog control efforts are within the BBP, LCP, and UPC.

The results from the willingness to pay question suggest Alabama landowners are primarily in favor of wild hog eradication. The results were heavily weighted towards those who preferred fewer/no wild pigs if present on their property, or if not present, preferred to keep it that way. Interestingly, of those who were willing to pay for wild hog eradication, a few landowners without wild hogs on their property were willing to pay the highest reported amounts. In the BBP where wild hog populations appear to be causing the most damage, landowners without wild hogs on their land were willing to pay nearly \$195 more per acre for eradication than those actually dealing with the animals on their property. This trend was also observed in survey responses from the AP and PP, though the differences in values reported were not as drastic as those for the BBP. In most cases it appears landowners without wild hogs on their land assumed they would cause more damage than they apparently do. The costs of wild hog introduction into an area may be felt over many years as their presence would create uncertainty and risk to a timber investment, perhaps explaining why these owners were willing to put such a premium on prevention.

Out of the nine respondents who felt they benefitted from wild hogs on their property, only one individual felt they needed to be compensated to allow wild hog eradication. This individual did not report making any income from wild hog hunting. Of the 51 respondents who reported making money from wildlife hunting leases on their property, only two individuals reported any profitability from hog hunting. The income reported from hunting leases specifically for wild hogs only constituted a small portion of the income produced from all hunting leases. These findings suggest the positive values respondents associate with wild hogs on their property cannot be explained by the

financial benefits alone. Around 56 percent of landowners who preferred the presence of wild hogs on their property indicated they hunted the animals; therefore, these landowners likely place a high value on having wild hogs available to hunt.

A few limitations should be acknowledged concerning the survey and its analysis. It is possible landowners did not have an accurate perception of damage done to their forest stand. The survey required respondents to recall a number of different figures and expenditures so any bias in answers could have been unintentional. Another area of improvement for this project concerns the growth and yield model used for the economic analysis. Although the model was created using the best resources available, it could potentially be improved to better reflect more accurate timber and pine straw harvest yields. The lack of a growth and yield model that accounts for a longer rotation given the management criteria specified for the stand level economic model proved to be a major obstacle that had to be overcome. Considering the goals of this project, we feel the model was sufficient in allowing us to look at how wild hog damage and resulting management decisions affected stand profitability.

During the course of the project it became evident that additional research was needed in a number of areas. The first is the need for a longleaf growth and yield model which can handle a variety of management options. This tool would prove valuable to anyone involved in planting or managing longleaf. The second is the need for more research on the potential for forestry insurance to cover wild hog damage. One survey respondent reported 95 percent seedling mortality on 410 acres of longleaf due to wild hog damage. In this case, replanting costs alone would be over \$84,000. Although this

was the largest case of damage reported, it illustrated the potential destructive ability of wild hogs in a forest plantation. Having an insurance policy to cover wild hog damage in longleaf plantations would prevent landowners from being forced to abandon their investment after unsustainable damage occurs. Third, additional research will be needed to investigate the efficiency of other control techniques as they become more common or available in the future. Wildlife Services is currently evaluating an aerial gunning program in Alabama which was initiated in 2016. Also, a wild hog toxicant is currently being researched for its potential use by Wildlife Services personnel. Adding to the variety of control techniques will be beneficial for wild hog management in the State, but time will tell if they are the solution.

In this study we sought to fill in the information gap concerning wild hog's impact on forest plantations. The results from this project offer State-wide damage estimates to forest stands where previously published estimates of this type did not exist. Additionally, the group's findings have identified regions in Alabama where reports of wild hog populations and damage appear to be more prevalent. Brook and van Beest (2014) acknowledge the importance of survey data in characterizing social-biological problems as well as providing a cost-effective and accurate means of acquiring data on a species' distribution. Likewise, as noted in Anderson et al. (2016), the self-reporting of wildlife damage to crops is common and provides reliable results. Having both a precise and broad understanding of how stand damage from wild hogs varies across tree species and region is valuable information for anyone involved in the forestry sector. Considerable effort is expended to manage wild hog populations by individuals, state, and federal agencies; the survey information from this study could help these entities

allocate resources to regions where wild hog problems are most severe. Additionally, the economic model and sensitivity analysis tables could prove beneficial to landowners and resources professionals who must make management decisions concerning forest stands amidst the growing threat from wild hogs. The information learned from this project will be a valuable resource for getting people thinking about the true cost of wild hog damage to forest stands as more longleaf is being planted across the Southeast.

CHAPTER 4

Wild Hog (*Sus scrofa*) Preference Among Planted Pine and Hardwood Seedlings and the Ecological Factors Influencing Young Forest Plantation Damage

4.1. Introduction

4.1.1. Current Status of Wild Hogs in the United States

Wild hogs (*Sus scrofa*) are among the most destructive introduced or exotic vertebrates to have become established in the Americas. As with their introduction to the continental United States (US), their recent expanse in range stems from anthropogenic causes. In 2000, wild hogs were known to occur in 21 states, but less than a decade later were found in at least 44 states (Mayer et al. 2000, Mayer 2009a). The rapid spread of this species across the US is likely due to a combination of illegal translocation or escaped animals from private or commercial fenced enclosures (Mayer 2009a). As a species, wild hogs quickly become established once introduced to an area due to a highly adaptable biology and prolific reproductive potential (Seward et al. 2004, West et al. 2009). Though problems stemming from the presence of wild hogs are not new to the US, in recent years more attention is being given to their rapid range expansion and the accumulating evidence of their deleterious impacts on the environment (Campbell and Long 2009, Sloodmaker et al. 2017).

4.1.2. *Wild Hog Damage in Relation to Forestry*

Wild hogs are particularly problematic to landowners because of their tendency to travel in groups and cause extensive damage to timberlands, pastures, and agriculture crops (Graves 1984, Seward et al. 2004, West et al. 2009). Damage to pine stands is likely more common in the Southeast where populations of wild hogs are dense and forested acres are abundant. In Alabama alone, timberland makes up nearly 70 percent of the State's total land area (AFC 2016). Wild hogs impact timber crops in a variety of ways including girdling trees through rubbing, damaging the lateral roots by rooting and chewing, and removing the bark of trees by tusking (Mayer 2009b). However, the most widespread and economically costly damage to the timber industry from wild hogs is the depredation of planted pine seedlings (Mayer 2009b).

Trees are most vulnerable to wild hogs during the initial years after planting or germination (Mayer 2009b, Sweeney et al. 2003). A single hog is reportedly capable of rooting up to six longleaf pine (*Pinus palustris* Mill.) seedlings a minute, destroying an estimated 400-1000 seedlings a day (Hopkins 1947, Wakeley 1954). These animals have the potential to cause complete crop failure in young timber plantations while seedlings are in their initial growth stages. In South Carolina, Lipscomb (1989) observed after two growing seasons, only eight of 845 longleaf seedlings survived hog predation in an unfenced area, versus 64 percent of seedlings surviving in fenced-in sites.

4.1.3. *Southeastern Tree Species Commonly Associated with Wild Hog Damage*

Wild hogs will occasionally cause damage to loblolly pine (*Pinus taeda*) and slash pine (*Pinus elliottii*), but the most extensive damage occurs with longleaf (Frost 1993, Wakeley 1954). Historical reports after a period of intensive planting by the US Forest Service in the mid-1900's documented entire longleaf plantations in Alabama, Florida, Louisiana, and Mississippi resulting in crop failure due to wild hogs (Lucas 1977, Wakeley 1954). Events like this could potentially become more commonplace as increasing amounts of longleaf acreage is planted by groups interested in longleaf restoration. Alabama has been one of the main southern states where longleaf restoration work has taken place in recent years (progress reports and other information about the Longleaf Pine Initiative is available online at https://www.nrcs.usda.gov/wps/portal/nrcs/detailfull/national/programs/initiatives/?cid=nrcsdev11_023913). Interestingly, a recent survey in Alabama of non-industrial private landowners reported more acres of longleaf were planted in 2013-2015 than loblolly (Chapter 3). It was also found that wild hogs damage more acres of planted longleaf than loblolly. Wild hog-induced crop failure would prove very problematic for states like Alabama where around 50 percent of timberlands are pine plantations (AFC 2016).

Longleaf pine is unique among southern pines in that they have evolved with landscapes exposed to frequent fire. While other tree species focus energy into rapid vertical growth during initial stages of development, longleaf may remain in a fire-resistant grass stage for several years before initiating vertical growth (Croker and Boyer 1975). This development trait, along with a few other adaptations (long needles, thick

bark, thickly scaled bud), allows longleaf to survive on landscapes with frequent fire intervals. During the grass stage, longleaf grows a thick tap root which may prove more appealing to wild hogs compared to root stems of other planted species (Mayer et al. 2000, Wood and Lynn 1977). Wood and Roark (1980) concluded that hogs were not actually consuming pine saplings, but instead were chewing on the roots to access the sap and starches then discarding the woody tissue. As a result of not actually ingesting the woody tissue, the group warned that wild hog's use of woody plant parts may be underestimated by stomach analyses.

Depredation to planted seedlings by wild hogs is not exclusive to southern pine species. Mayer et al. (2000) is the only study we are aware of to examine wild hog's impact on planted hardwood species. The group found wild hogs caused extensive damage to a number of planted hardwood seedlings in a wetland restoration area located in South Carolina. It was reported that of nine hardwood species planted, cherrybark oak (*Quercus pagodaefolia*), swamp chestnut oak (*Q. michauxii*), water hickory (*Carya aquatica*), and swamp tupelo (*Nyssa sylvatica* var. *biflora*) were the only species impacted by wild hog foraging activities. Non-effected seedling species included water oak (*Q. nigra*), green ash (*Fraxinus pennsylvanica*), persimmon (*Diospyros virginiana*), bald cypress (*Taxodium distichum*), and water tupelo (*Nyssa aquatica*). They postulated depredated tree species were more aromatic than non-impacted species which made them more appealing to the hog's highly developed sense of smell. It was also suggested wild hogs might have targeted the seedlings because the root tissue of nursery stock is often more succulent when transplanted than the corresponding root mass of natural seedlings. In a nursery, optimal growing conditions of nutrition and water management are

maintained to increase root volume prior to dormancy induction. As a result, seedlings with larger root volumes and increased starch and sugar levels are then transplanted to the field. They theorized the root tissue of nursery stock would be more succulent and appealing to wild hogs compared to natural seedlings. Additional findings from the study suggested one of the leading factors influencing seedling predation was the use of site preparation methods (e.g., prescribed burning) which enabled easy access to planted sites.

4.1.4. Ecological Factors Influencing Seedling Damage

A combination of ecological factors may determine the severity of wild hog depredation of planted seedlings (Mayer et al. 2000). Seedlings are found and removed through the rooting process, therefore factors affecting rooting will ultimately influence mortality from wild hogs. Rooting is the most widespread and observable type of damage done by wild hogs because all hogs root as a primary method of searching out food (e.g., roots, tubers, fungi and fossorial species) (Mayer 2009b). Wild hogs root throughout the year, but depending on location the intensity and frequency of rooting can be seasonal (Mayer 2009c). Ballari and Barrios-Garcia (2014) reviewed scientific literature pertaining to factors affecting food selection by wild hogs and found the use of food resource to be related to food availability, energy requirements, seasonal, and geographical variations. They also found some studies showing dietary differences between ages and sexes of wild hogs. In Schley and Roper's (2003) review of research concerning wild hog's diet, the group summarized that plant matter, both above and below ground parts, are a staple for the animal regardless of age, sex, or location. From the studies reviewed by Ballari and Barrios-Garcia (2014) and Schley and Roper (2003) it can be inferred seedling

predation is most likely to occur in the winter and spring than in the summer and fall based on the seasonal availability of food resources. In South Carolina, Wood and Roark (1980) suggested a similar timeframe for when longleaf seedlings are impacted by wild hogs; however, they concluded it is difficult to prove through stomach analysis because woody tissue is not actually ingested. In the Netherlands, naturally regenerated oak and beech seedlings were found to be selectively foraged by wild hogs in the late winter and spring (Bruinderink and Hazebroek 1996).

Site conditions in young forest plantations are widely varied, so determining specific conditions which attract wild hogs is difficult. Wild hogs are very selective in their choice of foraging areas which can be influenced by vegetative cover and/or soil moisture (Dexter 1998, Schley et al. 2008, Siemann et al. 2009, Wood and Roark 1980). Wild hogs may avoid pastures during abnormally dry years in favor for more hydric or mesic sites (Everitt and Alaniz 1980). Hunting pressure can also affect habitat usage by wild hogs. For example, in Alabama it was found wild hogs utilized wetland areas when hunting pressure was low but moved towards upland pine forests as hunting pressure intensified (Gaston et al. 2008). The degree of seedling predation in young forest plantations is site dependent and likely influenced by a combination of food availability, seedling accessibility, hog density, land cover, hunting pressure, and soil moisture.

4.1.5. Research Needs and Project Description

There is an apparent lack of research pertaining to wild hog behavior in young forest plantations. Most reports and scientific literature on wild hog predation of seedlings concern naturally regenerated pine, oak, and beech species (Bruinderink and

Hazebroek 1996, Hanson and Karstad 1959, Ickes et al. 2005, Lipscomb 1989, Siemann et al. 2009, Sweitzer and Van Vuren 2002). Few reports or research exist which specifically address wild hog depredation of planted seedlings. Mayer et al. (2000) is the only study we are aware of which tests for wild hog preference among planted hardwood seedlings and reported factors which influenced observed damage. With forestry trends in the Southeast indicating an increase in longleaf pine planting, any additional research pertaining to plantation damage by wild hogs could prove valuable to landowners and resources professionals.

The goal of this project was to build on the previous works of Mayer et al. (2000) and examine wild hog preferences between planted pine and hardwood seedlings. It was also of interest to observe the ecological factors which might influence differing amounts of wild hog damage in young forest plantations. We conducted a field study at two research locations with uniquely different site conditions in order to test for wild hog preference among planted seedlings and make inferences about the damage. Such information is beneficial in guiding forest management decisions as the threat from wild hogs becomes more widespread.

4.2. Methods

A number of tasks were created in order to guide this project through the research process. These tasks were designed to help accomplish the goal of determining if wild hogs had a preference among selected seedling species and identifying ecological factors that may influence forest plantation damage. Three main tasks were associated with this portion of the study:

Task 1. Identify two study sites within reasonable proximity to a creek drainage system with evidence of wild hogs in the area. Collect data on soil type, precipitation, and landcover to describe site conditions.

Task 2. Conduct an observational study to identify if wild hogs have a preference among planted seedling species.

Task 3. Establish camera traps to detect wild hog presence in the research area throughout the study period.

4.2.1. *Site Selection and Description*

In order to accomplish Task 1, two suitable sites were identified on private properties in Bullock County, Alabama. The first site (S1) (N 32° 10' 1.999", W 85° 37' 52.32") was located on the Auburn University Turnipseed-Ikenberry Place approximately 10 miles from the second site (S2). The location of the second site has been omitted at the request of the landowner.

Areas chosen for the study sites were less than 100 yards from a creek drainage system. Drainage systems are often utilized by wild hogs for cover and ease of movement; therefore, placing research sites close to drainage systems would reduce the amount of time it would take the animals to find the planted seedlings. Reducing the search time for wild hogs to find seedlings was desirable for this project given the year-long study period.

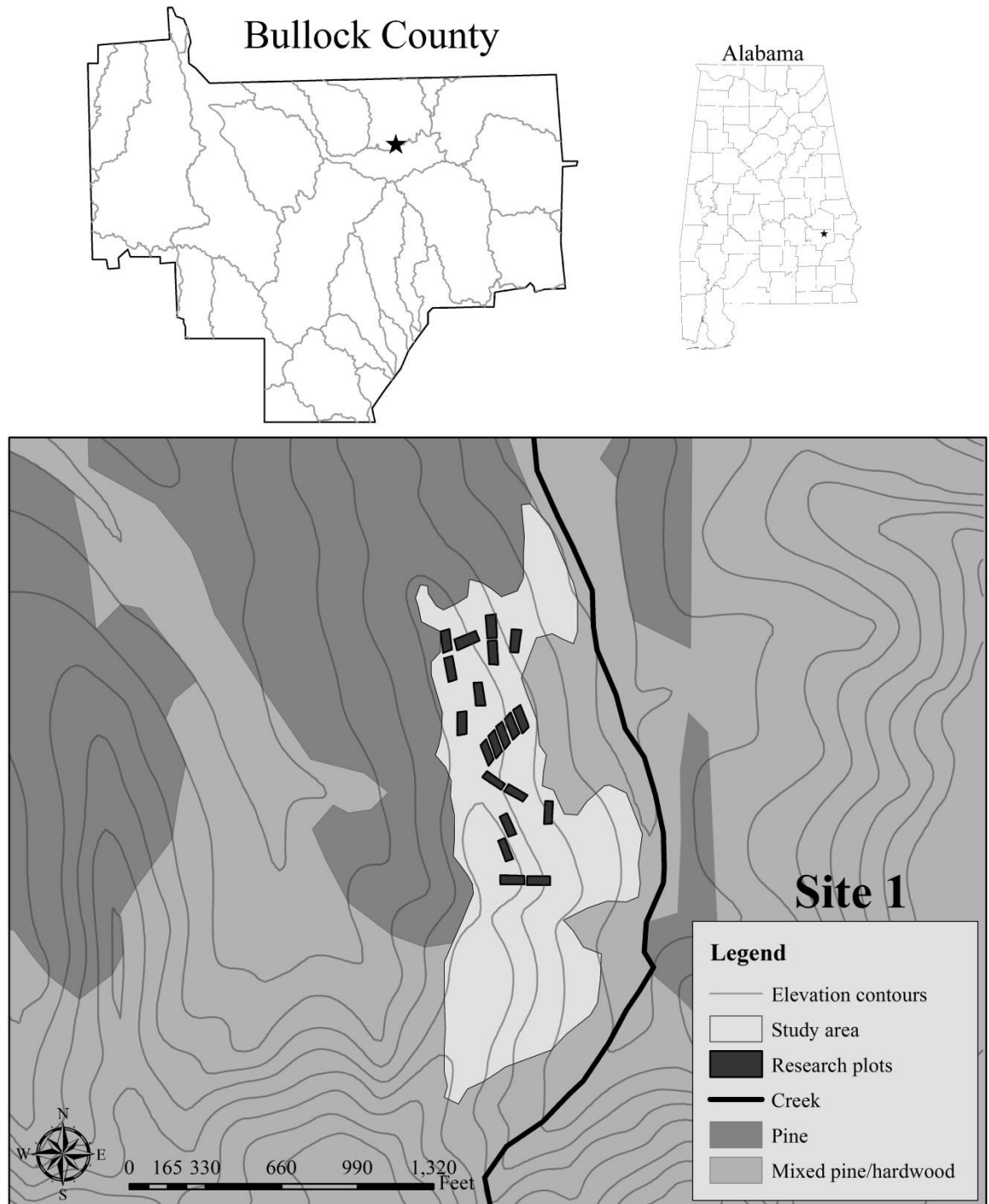
Wild hog populations were confirmed at S1 by setting up a bait station consisting of soured corn and a trail camera (Moultrie M-1100i Mini Game Camera) to capture visitation. Signs of wild hog presence in the field and along the drainage system was identified by rooting, tree rubs, and tracks. There was very little, if any, hunting pressure at S1. The groundskeeper for the property did not know of any previous attempts to control wild hogs aside from the occasional shooting on sight by turkey and deer hunters. A handicap shooting house was present on the southern end of the field, but it is unlikely wild hogs were harassed at the other end of the field given the contour of the land.

The 28.5 acre field where S1 (Figure 4.1) was located had previously been a pecan (*Carya illinoensis*) orchard and still retained a number of pecan trees dispersed throughout the field. The landcover in the field was primarily dominated by bahaigrass (*Paspalum nontatum*). Bahaigrass is a desirable cover species in pecan orchards because it can be easily mowed and is relatively shade tolerant. This species is a deep-rooted perennial which grows low to the ground and spreads with stolons and stout, scaly rhizomes to form dense mats (Houck 2009). Bahaigrass is resistant to drought and well suited for southern pastures. In September 2017, the field was mowed around the plots to

allow easier access for observational visits. Other vegetation observed growing among bahagrass was blackberry (*Rubus* spp.), bushy bluestem (*Andropogon glomeratus*), Sericea lespedeza (*Lespedeza cuneata*), late boneset (*Eupatorium serotinum*), and purple tridens (*Tridens flavus*).

The surrounding forest type varied from pine to mixed hardwood species. The pine stands were located at the top of the hill in the northern section of the study area while the mixed pine and hardwood forest constituted the southern, bottomland portion. Pine species included loblolly and shortleaf pine (*Pinus echinata*). The overstory in the bottomland portion was dominated by loblolly and water oak while the mid-story consisted of American sycamore (*Plantanus occidentalis*), eastern red cedar (*Juniperus virginiana*), red mulberry (*Morus rubra*), and sweetgum (*Liquidambar styraciflua*). The understory was comprised of Chinese privet (*Ligustrum sinense*), greenbrier (*Smilax* spp.), Japanese honeysuckle (*Lonicera japonica*), Mary's grass (*Microstegium vimineum*), and spike uniola (*Chasmanthium laxum*).

Figure 4.1. Location of research area at site 1 (S1) for the planted seedling preference by wild hogs study in Bullock County, Alabama (March 2016-March 2017).



The majority of the field at S1 was located on a hill side with an estimated slope of 5 to 20 percent. Taxonomic information regarding the soil series at S1 was collected using the United States Department of Agriculture's (USDA) Web Soil Survey (Available online at <https://websoilsurvey.sc.egov.usda.gov/App/HomePage.htm>). The soil in the field at S1 had properties consistent with Conecuh series soils (Soil Survey Staff 2014a). They are formed in clayey and shaley marine sediments. The family classification for the Conecuh series is fine, smectitic, thermic, Vertic Hapludults. Conecuh series soils are found on uplands and hill slopes in the Southern Coastal Plains. They are a moderately well drained soil with very slow permeability due to the higher clay content in subsurface horizons. These soils have a surface horizon consisting of sandy loam (0-5 inches thick) and Argillic subsurface horizons (5-39 inches thick). Argillic horizons are primarily composed of alluvial clays and indicative of increasing clay content with increasing depth. Soils at the summit of the hill were more eroded than those found below.

Soils at the bottom of the hill and extending to the drainage system were dominated by Mantachie series soils (Soil Survey Staff 2013c). The family classification for the Mantachie series is fine-loamy, siliceous, active, acid, thermic Fluventic Endoaquepts. Mantachie series soils are formed in loamy alluvium and are somewhat poorly drained with moderate permeability. These soils are commonly associated with flood plains in the Southern Coastal Plains. The surface horizon is made up of fine sandy loam all the way through the A-horizon (0-11 inches thick). Signs of gleying are found below the A-horizon which is indicative of aquic soil conditions (19+ inches below).

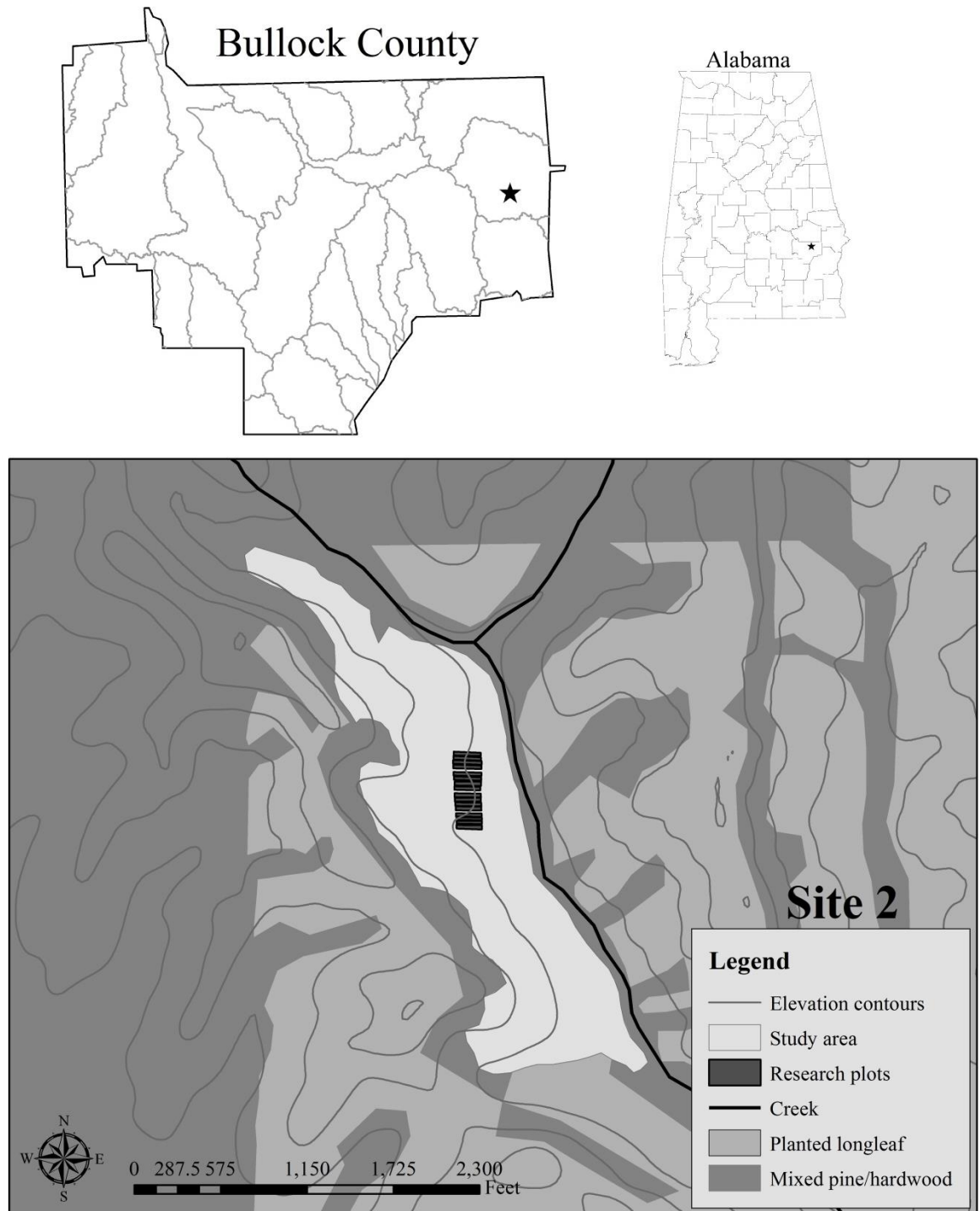
S2 (Figure 4.2) was located on 90 acres of cutover land previously planted with longleaf pine in January 2015. The longleaf plantation sustained such high levels of wild hog damage over the pursuing weeks that the landowner was forced to replant the following year. The landowner at S2 described frequent sightings of wild hogs in the field of interest. Evidence of the landowner's reports was supported by signs of wild hog presence, so bait stations were not necessary as at S1. The landowner utilized hunting and trapping in an effort to decrease the wild hog population on the property in and around the area where the study site was located. The field containing the study site was part of a large acreage primarily managed for game species and longleaf production.

The field at S2 was a cutover site with frequent stumps and woody debris left over from a loblolly stand clearcut in 2010. After the clearcut, the remaining tree cover available to wildlife immediately surrounding the research area was mixed pine and hardwood forests along the stream management zone (SMZ). Tree species found in the SMZ primarily consisted of loblolly, laurel oak (*Q. laurifolia*), and water oak. The most commonly found vegetation in the understory of the SMZ was river cane (*Arundinaria gigantea*) and spike uniola interspersed with beautyberry (*Callicarpa americana*) and muscadine (*Vitis rotundifolia*). Over the peak of the hill, the adjacent forest to the east of the research site consisted of mixed pine and hardwood forest including laurel oak, loblolly, post oak (*Q. stellata*), shortleaf pine, southern red oak (*Q. falcata*), sweetgum, and water oak.

A prescribed burn was performed by the landowner before the initial planting of longleaf in September 2015. Bare-ground was visible during the initial months of the

field study but became less common as panicgrass (*Panicum* spp.) became more abundant. Other species frequently observed were common ragweed (*Ambrosia artemisiifolia*), dogfennel (*Eupatorium capillifolium*), purple false foxglove (*Agalinis purpurea*), late boneset, and sericia lespedeza.

Figure 4.2. Location of research area at site 2 (S2) for the planted seedling preference by wild hogs study in Bullock County, Alabama (March 2016-March 2017).



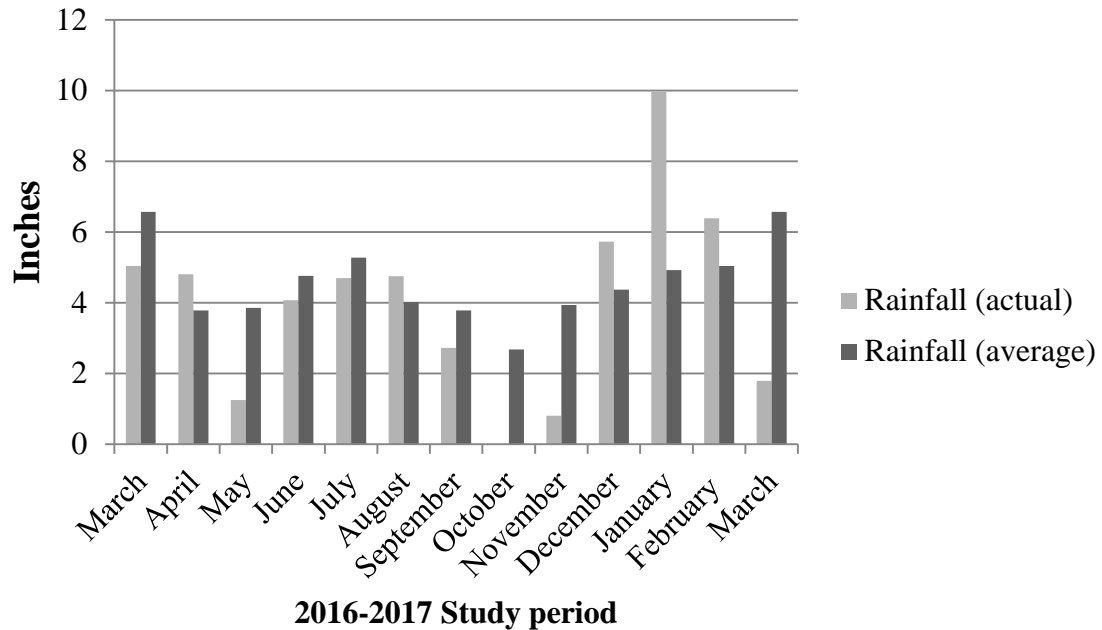
S2 was located on hill side with an estimated slope ranging from 5 to 15 percent. Taxonomic information regarding the soil series at S2 was also collected using the USDA's Web Soil Survey. The soil in the field had properties consistent with Luverne series soils (Soil Survey Staff 2014b). They are formed in stratified marine sediments. The family classification is Fine, mixed, semiactive, thermic Typic Hapludults. Similar to Conecuh series, Luverne series soils are found on uplands in the Southern Coastal Plains but in areas where erosion has caused the landscape to become dissected. One of the more prominent differences between the Luverne and the Conecuh series soils is in their drainage and permeability properties. Luverne series soils are well drained with moderately slow permeability due to the E-horizon underlying the epipedon. These soils have a surface horizon consisting of fine sandy loam (0-7 inches thick) and Argillic subsurface horizons (7-30 inches thick). In contrast to S1, S2 was associated with higher erodibility and slightly increased amounts of sand and silt.

Soils at the bottom of the hill and extending to the drainage system were dominated by Blanton (Soil Survey Staff 2013a) and Bonifay (Soil Survey Staff 2013b) series soils. The family classification for the Blanton series is Loamy, siliceous, semiactive, thermic Grossarenic Paleudults. The family classification for the Bonifay series is Loamy, siliceous, subactive, thermic Grossarenic Plinthic Paleudults. Blanton and Bonifay series soils are formed in sandy and loamy marine deposits. They are excessively to moderately well drained with moderate to slow permeability. Blanton series soils are commonly associated with upland and stream terraces in the Southern Coastal Plains while Bonifay series soils are found on ridges and side slopes. Blanton series soils have thick layers of fine sand above the underlying Argillic horizons (0-52

inches thick). Similarly, the Bonifay series has a thick layer of sand above the Argillic horizon (0-57 inches thick). This area was commonly flooded except during periods of drought.

Monthly precipitation data during the course of the field study is presented in Figure 4.3. These data were collected from the US climate data website (Available online at <https://usclimatedata.com/>). Also included in the figure is the normal monthly precipitation data based on historic averages for comparison purposes. In 2016, severe drought conditions were observed throughout Alabama during October and November. Research sites did not receive any rainfall for a 62 day period from September 28 to November 29, 2016. The overall amount of precipitation for the study period was 12 percent lower than normal. The overall rainfall for the duration of the study period was 52.03 inches.

Figure 4.3. Observed and expected monthly precipitation data for Bullock County, Alabama from March 2016 to March 2017.



4.2.2. Description of Experimental Design for Testing Planted Seedling Preference by Wild Hogs

For Task 2, seedlings were purchased from nurseries in Alabama, Georgia, and Tennessee. Seedling species used in this study included longleaf pine, loblolly pine, cherrybark oak, chinkapin oak (*Q. muehlenbergii*), and persimmon. The only seedlings not bareroot were longleaf, which were containerized. Species chosen for this study had two or more of the following qualifications: 1) association with wild hog damage, 2) commonly planted in the Coastal Plains region, and 3) availability from nurseries. Longleaf, loblolly, and cherrybark oak are some of the seedling species most often associated with wild hog damage. In chapter 3, longleaf and loblolly pine plantations sustained the highest levels of wild hog damage according to a survey of non-industrial

private landowners in Alabama. Cherrybark oak was found to be the most highly damaged planted hardwood species in a wetland restoration area in South Carolina (Mayer et al. 2000). Chinkapin oak and persimmon can be commonly found growing in the Coastal Plains and were readily available from nurseries.

Planting procedures took place in March 2016. Each site was divided into 4 blocks. The blocks at S1 were further divided into five, one-tenth-acre plots while the blocks at S2 were divided into four, one-tenth acre plots. Plots at S1 were oriented so they were not shaded by pecan trees located occasionally throughout the field. There were no trees in the field at S2 so plots were oriented sequentially. Each plot within a block was assigned a tree species to be planted through random assignment without replacement. Plots were planted with the equivalent of 545 trees per acre with 8 x 10 foot spacing between trees. Each tree was assigned a numbered flag placed next to the seedling. Ten seedlings in each plot were randomly selected to serve as the control and received protective netted-tubes. The tree tubes were anchored by bamboo or wooden stakes and secured with zip-ties. Planting procedures were uniform between the two sites with the exception of the longleaf seedlings at S2. In February 2016, the landowner planted 400 acres of longleaf which included the cutover land the research site was located on. Due to this previous planting, it was not necessary to plant additional longleaf. Within each one-tenth-acre plot at S2, hardwood and loblolly seedlings were planted between longleaf seedlings. Since every other seedling at S2 was longleaf, a plot from each block was chosen through random assignment without replacement to serve as the longleaf plot for control samples and measurements.

After planting, the ground-line diameter (GLD) was recorded for every fifth seedling in a plot as well the seedlings serving as the control. This measurement was repeated the following year at the conclusion of the study. The status of seedlings was monitored throughout the experiment with monthly visits, except for months corresponding with deer and turkey hunting seasons when the sites were not accessible. During monthly visits, each seedling's status would be marked as either 'alive', 'dead', or 'hogMortality'.

4.2.3. Camera Trap Deployment for Wild Hog Detection

In order to fulfill Task 3, three camera traps (Moultrie M-1100i Mini Game Camera) were set up at each site on game trails between the drainage system and the planted seedlings. Passive observation techniques were chosen to avoid potential affects from using bait sites. Using bait sites to estimate wild hog density might have altered foraging behavior or attracted unwanted attention from hunters. Cameras were set to take one photo every five minutes and had trigger sensitivity on the high setting. The cameras were kept operational throughout the majority of the study period except for a short period of time when camera maintenance was performed (between July and August 2016). The quantity of pictures featuring wild hogs collected each month were weighted based on the number of cameras at each site and the number of days cameras were operational. This was done to compensate for when cameras were not in use for maintenance or technical malfunctions.

4.2.4. Statistical Analyses for Field Study Data

Statistical analyses were performed using SPSS (IBM Corp. 2015) to test the following hypothesis: 1) seedling mortality due to wild hogs is different among planted species. Two additional hypotheses were created because of a severe drought during the field study. These hypotheses were 2) mortality rates among species were different, and 3) the seedlings in the mortality group would have larger GLD measurements than the seedlings that survived.

A Pearson's Chi-square test was used to test hypothesis 1 to determine if the number of seedlings associated with wild hog damage was significantly different among seedlings species. A Pearson's Chi-square test was also used to test hypothesis 2 in order to determine if the non-hog related mortality was different among species. Lastly, a one-way ANOVA was performed to see if the initial GLD for seedlings in the mortality group was significantly larger than those that survived.

4.3. Results

4.3.1. *Results for the Wild Hog Seedling Preference Experiment and Statistical Analyses*

Table 4.1 summarizes the results of the descriptive statistics performed on the seedling data. At the conclusion of the study seedlings either survived (Alive), suffered mortality not caused by wild hogs (Dead), or were destroyed by wild hogs (HogMortality). Mortality from wild hogs was easily distinguished by observing rooting where the seedling had been originally planted. In some cases the seedling was found nearby with the root stock having been masticated. When seedling mortality occurred from wild hogs, the time of year and location of the damage with respect to hill slope was recorded. At S1, 89 percent of wild hog damage to seedlings occurred in the spring and 11 percent occurred in the summer. All seedling mortality from wild hogs at S1 were from plots located at the bottom of the hill. At S2, 74 percent of wild hog damage to seedlings occurred in the spring and 26 percent occurred in the summer. Seedling damage at S2 was evenly distributed between the top, middle, bottom of the hill.

The percent of seedling mortality due to wild hogs was low at S1. Cherrybark had the most seedlings damaged (n=5), followed by longleaf (n=2) and loblolly (n=2). There was no hog-related mortality in chinkapin or persimmon. Chi-square analyses were used to determine if observed frequencies of mortality differed from that which would be expected. The expected mortality is the number of seedlings that would have had to have suffered mortality due to wild hogs if the damage was equally distributed. The Chi-square analysis used to test if hog-related seedling mortality was different among planted species (Hypothesis 1) at S1 was not significant for hog-related mortality (n=9) compared to

other mortality (n=243) ($\chi^2 = 2.01$, $d.f. = 2$, $p > 0.05$), while the Chi-square analysis used to test hypothesis 1 for S2 did detect a significant difference in hog-related mortality (n=96) compared to other mortality (n=626) ($\chi^2 = 75.34$, $d.f. = 4$, $p < 0.001$). At S2, longleaf was the most heavily damaged by wild hogs (n=77) and had more than double the expected frequency of mortality (n=39.8). Cherrybark was the second most damaged species by wild hogs and had similar observed mortality (n=10) as expected mortality (n=11). Persimmon seedlings had around half the observed wild hog mortality (n=5) than the expected mortality (n=10.6). Loblolly had considerably less observed mortality (n=3) than the expected mortality (n=12.4). Lastly, chinkapin had the lowest number of seedlings damaged (n=1) compared to the mortality expected (n=22.2).

Alabama experienced a dry year in 2016 with a two month drought which started in October and lasted through November which resulted in lower survival of seedlings than would normally be expected. Non-hog related mortality is summarized for each species in Table 4.1. The Chi-square analysis used to test if mortality rates among species was different (Hypothesis 2) indicated a significant difference existed between the expected and observed survival and mortality of seedlings species (S1: $\chi^2 = 94.12$, $d.f. = 4$, $p < 0.001$; S2: $\chi^2 = 173.58$, $d.f. = 4$, $p < 0.001$). Based on the frequency of observations the Chi-square analyses indicated the expected mortality at S1 and S2 for seedlings species was 70.2 and 84.4, respectively, with the exception of longleaf at S2 which had an expected mortality of 307.5 due to the larger sample size. At S1, loblolly (n=114) and cherrybark (n=96) experienced higher mortality than the expected frequency. Longleaf (n=42) and persimmon (n=38) performed the best despite severe drought conditions and had mortality less than half the expected frequency. At S2, longleaf (n=222) again had

the lowest amount of mortality compared to the expected frequency. Loblolly (n=93) mortality was slightly higher than the expected frequency while cherrybark (n=83) and persimmon (n=80) had observed amounts of mortality just below the expected frequency. At 77 percent mortality, chinkapin (n=167) suffered the largest losses.

Descriptive results of the average GLD measurement taken at the beginning of the study are presented in Table 4.1. The GLD averages for each species are organized by status. For the status titled HogMort, GLD averages or standard deviations were not available for all species because the sample group for measured seedlings damaged by wild hogs only contained a single sample or none. Due to the low number of measured seedlings being damaged by wild hogs, it was not possible to statistically test for a relationship between seedling size and hog damage. However, a general trend was observed in that smaller seedlings (indicated by a smaller GLD) survived better compared to larger seedlings. In order to test this relationship, a one-way ANOVA was conducted for each species at each site to see if the average initial GLD of surviving seedlings was statistically different from those in the mortality group (Hypothesis 3). The results of the analyses were only significant at S2 for chinkapin oak, $F(1,74) = 6.50, p < 0.05, \eta^2 = 0.08$, and persimmon, $F(1,73) = 9.48, p < 0.05, \eta^2 = 0.12$. Based on the results of the analyses, we fail to reject the null hypothesis for chinkapin and persimmon sample groups at S2. These results suggest the initial GLD for chinkapin and persimmon were significantly larger in seedlings that died than in seedlings that survived. The one-way ANOVA results for chinkapin oak are consistent with field observations made when planting S2. Chinkapin seedlings were larger and more cumbersome to plant compared to other species.

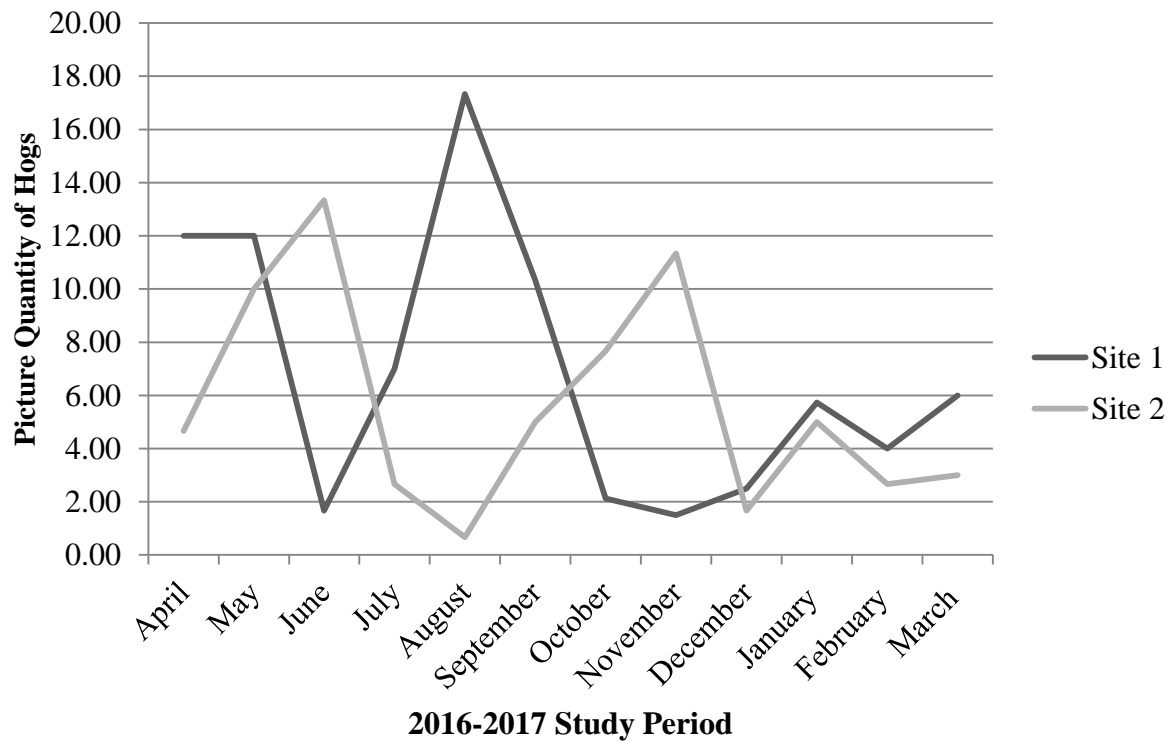
Table 4.1. Wild hog seedling preference results including percent survival, mortality, and wild hog mortality with respective averages for ground line diameters (GLD) and standard deviation (SD) for sites in Bullock County, Alabama.

Site	Species	% Alive	Mean GLD	SD	% Dead	Mean GLD	SD	% HogMort	Mean GLD	SD
S 1	Loblolly	48%	3.63	0.92	51%	3.87	0.89	2%	4.74	2.18
	Persimmon	83%	5.86	1.63	17%	6.23	1.78	0%	-	-
	Chinkapin	72%	6.08	1.94	28%	7.62	3.79	0%	-	-
	Cherrybark	56%	7.63	2.17	41%	8.23	3.56	5%	6.19	-
	Longleaf	81%	11.31	2.44	18%	12.85	3.79	5%	-	-
S 2	Loblolly	57%	4.01	0.68	42%	4.16	1.15	3%	4.60	-
	Persimmon	63%	6.56	1.55	35%	8.00	2.34	6%	5.32	-
	Chinkapin	23%	10.84	4.21	77%	13.64	4.82	1%	-	-
	Cherrybark	62%	7.79	2.37	34%	8.24	2.88	12%	9.37	2.51
	Longleaf	65%	10.67	2.28	26%	9.42	3.21	26%	12.98	1.53

4.3.2. Results of Wild Hog Detection in Research Areas

The results of wild hog detection in the research area for each site are presented in Figure 4.4. Wild hogs frequented the research areas more during the spring and summer months compared to winter months. The cyclical nature of the results for S2 was most likely due to intense periods of hunting and trapping efforts by the landowner. The largest drop in wild hog detection around S2 in August can be explained by the landowner hunting over the study site more frequently compared to other months. There was another large drop in detection at the first site during the month of June, what caused the hogs to disappear from the research area during this time is unknown.

Figure 4.4. Wild hog detection data collected at site 1 (S1) and site 2 (S2) study areas in Bullock County, Alabama. Y-axis values are weighted based on the number of days cameras were operational.



In a similar study conducted in South Carolina, Lipscomb (1989) used the following density indices to describe the population of wild hogs in an area: low (less than one hog per 20 acres), medium (more than one hog per 20 acres but less than 1 per 10 acres), and high (more than 1 hog per 10 acres). Our detection results suggest the wild hog population at S1 and S2 would qualify in the high density range.

There were differences between sites for the time of year when piglets were observed in picture data. The number of months piglets appeared in picture data was higher for S2 than S1. At S1, young piglets were seen in April and May 2016 and February 2017. At S2, piglets were observed in May, June, July, September, and

November 2016 and January 2017. These results suggest offspring were born in the late winter and spring at S1 while offspring were born in each season at S2.

4.4. Discussion

4.4.1. Seedling Preference Results in Relation to Other Research

The main goal of this project was to determine if wild hogs had a preference among planted pine and hardwood seedlings or if damage was due more to chance encounter. Although Mayer et al. (2000) showed wild hog's had a preference among planted hardwood species, there are no previous studies which examine both planted pine and hardwood species.

The results at S2 were consistent with historic and scientific reports of wild hogs preference towards planted longleaf pine and cherrybark oak. If the resulting wild hog damage to longleaf and cherrybark seedlings from S2 were extrapolated to a per-acre bases, the result would be 48 and 25 seedlings per acre, respectively. The per acre seedling loss may be more substantial in situations where cherrybark is planted because planting densities are normally not as high as in longleaf plantations. Longleaf appeared to be the most highly preferred seedling species among those tested. It should be noted that at S2 the sample group for longleaf was nearly four times larger than the other species tested; therefore, there was a higher chance of longleaf being damaged. Only a small portion of the total area of the property planted with longleaf was monitored for the purposes of this study, so similar damage could be assumed to be occurring elsewhere in unmonitored areas.

4.4.2. *Site 2: A Case Study*

The events which occurred at S2 in 2015 serve as an important case study of the potential devastating impacts wild hogs can have on young longleaf pine plantations. According to the landowner, 400 acres of longleaf were planted in winter of 2015, and one month later wild hogs had destroyed around 95 percent of seedlings. After investing considerable time and money managing the wild hog population on the property, the landowner was able to substantially reduce hog-related mortality but did not eliminate it all together. During the year-long duration of the study, the landowner reportedly killed around 300 wild hogs on the property while spending an estimated \$5,000 on hunting costs and trapping material. This case study illustrates how, if left unmanaged, wild hogs can cause substantial financial loss for forest plantation owners. Through a relatively small investment in management efforts the damage can be mitigated. Schely et al. (2008) postulated that wild hog density was the most important predictor of damage and reducing the population would be a useful management tool. Population reduction through trapping and hunting appears to have been an efficient way for the landowner to reduce damage in the forest plantation compared to the previous year when no control was implemented.

Another observation worth noting in relation to mitigating damage to seedlings, is that the screened tree tubes were 99.5 percent effective at protecting seedlings in the control group from wild hog damage. Only two seedlings in the control group protected by plastic tubes were destroyed by hogs. Perhaps the tubes with wooden or bamboo stakes were a physical deterrent which discourage wild hogs away towards more easily

accessed foods. More research would be needed to test tree tube effectiveness in preventing wild hog damage. Tree tubes or some related type of individual tree protection could be an inexpensive way to add extra protection to planted seedlings.

4.4.3. Potential Explanations for Trends in Wild Hog Detectability

It was interesting to note how wild hog detectability results varied between sites with respect to the differing amounts of hunting pressure. S1 received low hunting pressure while S2 had high hunting pressure. In May 2016, the observational visit to S2 attracted the attention of the landowner who took notice at the amount of rooting which had occurred the previous months. It was noted how the landowner made plans to start hunting over the field site which could explain why wild hog detection dropped between the middle of July to the middle of August. After a period of time not seeing wild hogs in the field, the landowner likely concentrated hunting and trapping efforts elsewhere on the property. The frequent removal of hogs by the landowner could have been the cause of the increased amount of recruitment observed throughout the year at S2 in contrast to S1 where piglets appeared in picture data primarily in the spring.

When natural sources of food are abundant, wild hogs may not be attracted to bait and trapping can be inefficient (West et al. 2009). The landowner at S2 indicated trapping efforts became more intense in the winter after hard mast became less available. Trapping during this time reportedly yielded better results because hogs were more likely to enter traps in search of food. The second drop in December could be explained by the landowner having more success with trapping after the sources of mast were exhausted and wild hogs were easier to trap. These conclusions are supported by Barrett and

Birmingham (1994) who documented that trapping success was low during periods of heavy acorn production.

Gaston et al. (2008) found varying levels of hunting pressure caused modifications in wild hog behavior. It is possible that the difference in hunting pressure differences between sites caused a variation in wild hog behavior as well as seedling damage. Optimal foraging theory suggests time spent foraging in the open is a trade-off between accessing optimal food sources and the risk associated with leaving cover. Under the premise of the optimal foraging theory, wild hogs at S2 would likely have lingered in the open for less time and been forced to forage more quickly than those at S1. In contrast to S2, wild hogs at S1 would be able to forage in the open for longer periods of time and be more selective. These wild hogs might not have found seedlings to be the most desirable food source in the pasture and therefore avoided them.

4.4.4. Seedling Depredation: A Learned Behavior

Through studying the damage in the experimental plots, insights were gained on factors potentially influencing wild hog damage in young forest plantations. These insights are beneficial to landowners and managers and could help prevent wild hogs from causing heavy financial losses among forest plantations. Perhaps the most important postulation derived from the evident difference in levels of seedling damage between the two sites is that seedling predation of planted species is apparently a learned behavior among wild hogs.

As habitat generalist, wild hogs are an efficient invasive species because they are able to meet their dietary needs even in non-native ranges (Ballari and Barrios-Garcia 2014). This is accomplished by using heightened olfactory senses to explore and discover desirable foods (Moulton 1967). If all wild hogs found longleaf as a highly favored food source then the hog-related mortality at S1 would have been much higher. The amount of rooting next to one of the plots confirmed wild hogs had found the longleaf seedlings, yet only 1 percent of all available longleaf seedlings were consumed. Additionally, extensive rooting was done in 80 percent of one cherrybark plot at S1, yet no seedlings were damaged or consumed.

In contrast to the wild hog population at S1, the wild hogs at S2 would have been less naïve about planted seedlings as a food source because the landowner had multiple-aged stands located on the property which meant planted seedlings had been available in previous years. It is possible a few of the remaining hogs still remained on the property were familiar with planted longleaf seedlings as a food source. Predation on seedlings could have been observed by other hogs which would explain why seedling damage was more common at S2. This theory has important management implications for forest plantation owners because if wild hogs begin to learn that planted seedlings are a desirable food source then that population would need to be removed so the behavior would not be passed along to other hogs. On the other hand, if the population of hogs is naïve to eating the planted seedlings it may prove beneficial to leave them alone; otherwise, new wild hogs that recognize the seedlings as a desirable food source may move into the area. This concept concerning wild hogs and planted seedlings has not been encountered in scientific literature and warrants more research.

4.4.5. *The Impacts of Land Cover and Soil Type on Wild Hog Damage*

Another observation with important management implications for forestland owners is that cutover sites appear to create very attractive foraging areas for wild hogs. The woody debris left after the clearcut at S2 appeared to be an attractant to wild hogs interested in searching for invertebrates among the decomposing logs and stumps. Similarly, a study in South Carolina by Zengel and Conner (2008) found a positive association between rooting frequency and amounts of coarse woody debris. Invertebrates make up a small percentage of wild hog's diet but play an important role as a source of protein required year round (Ballaria and Barrios-Garcia 2014, Schley and Roper 2003, Wood and Roark 1980). Consequently, wild hogs were frequently attracted to foraging in the research site and were more likely to encounter seedlings in their search for food.

Interestingly, seedling damage at S2 was minimal later in the year as seedlings were hidden under thick amounts of vegetation which would have made them difficult to access. The idea that depredation of planted seedlings is related to accessibility was suggested by Mayer et al. (2000) who found that areas pretreated (clearing and burning) were the most severely impacted by hogs. The conditions at S2 would support this observation but not at S1. In September 2016, the bahagrass and briars had made observational visits difficult, so the field was bushhogged around the seedling plots. After the grass had been mowed the seedlings would have been very accessible to hogs for the rest of the study period, yet seedling damage did not occur. This suggests that factors

other than accessibility may be more important in influencing predation of planted seedlings by hogs.

Wild hog damage at S1 was notably less than the amount that occurred at S2. Given the differences in the two sites it is likely land cover played a role in the observed damage. Initially it was assumed the damage at S1 would have occurred at a higher level than observed because wild hog damage in pastures is quite common. Schley et al. (2008) found wild hog damage to grasslands in Germany to be severe and occur frequently. Additionally, the 2016 survey of non-industrial private landowner in Alabama (Chapter 3) indicated that grassland/pasturelands was one of the top three most commonly reported land cover types for wild hogs sightings. Damage was mostly limited to a small section in the northeastern quadrant at the bottom of the hill where water drainage would have kept the soil more moist compared to the rest of the pasture. Everitt and Alaniz (1980) observed wild hogs avoided pastures in abnormally dry years which may explain why damage at S1 was minimal.

The Conecuh series soil appeared to be an important factor in limiting hog damage to periods when rain events allowed hogs to penetrate the soil surface in search of food. The high content of clay, which causes this particular soil to have very slow water permeability, created a cement-like barrier when soil was devoid of sufficient amounts of moisture. The Luverne and Blanton series soils at S2 also had Argillic subsurface horizons but had thicker A and E horizons separating the soil surface from the layer where increases in clay are found. With 2016 being a drier year than normal in Alabama, it is possible the field was not conducive for wild hogs to root and forage for

food compared to other areas. It was evident the majority of wild hog activity was concentrated in areas near the drainage system where the soils transitioned from the Conecuh series to the Mantachie series. The higher sand and silt content of the Mantachie series soil would have made rooting relatively easier.

As with S1, soil type was also a factor affecting hog damage at S2. The Blanton, Bonifay, and Luverne series soils appeared to be more beneficial to rooting activities because wild hogs will root and dig to depths of 24 inches below the soil surface in search of food (Schley and Roper 2003). Despite how dry it was throughout the year, the sand and loam components of the soil at S2 made it easier to penetrate the soil surface than at S1. The friability of the soil structure was also evident because of the amount of erosion occurring in the area.

4.4.6. Changes in Seedling Damage in Respect to Food Availability

The vegetative diversity was greater at S2 and included a large abundance of panicgrass which is one of the most frequently consumed herbages by wild hogs in this part of the world (Wood and Roark 1980). Panicgrass and other flora would mainly have been consumed during the spring when new shoots and herbs were most luxuriant (Ballari and Barrios-Garcia 2014, Wood and Roark 1980). As was initially expected, the majority of seedling damage occurred in the spring of 2016 at both sites as hogs foraged for succulent shoots and roots. The amount of seedling damage decreased by a third during summer months compared to the spring which was not surprising considering wild hog's summer diet consists primarily of fruit (Ballari and Barrios-Garcia 2014). Seedling damage was minimal through the fall and winter months as hard mast became available.

Seedling damage was expected to be higher in the winter as above-ground plant parts became scarce and sources of hard mast were depleted; however, this was not the case at either site for reasons that could not be determined.

4.4.7. Impact of Drought Conditions and Areas of Project Improvement

The outcome of this study was impacted by the unusually dry year for Alabama as well as the drought in October and November. We believe this caused the high mortality rates observed in seedlings and potentially reduced the amount of damage from wild hogs both to the sites and the seedlings. Casperson and Kobe (2001) showed soil moisture and tree species' ecological response to environmental stress have a dramatic effect on seedling survivorship. They suggested seedling size might have played a role in the differences in mortality rates among species tested. Conifer sapling mortality did not vary with respect to soil moisture due to the needle leaves having low surface area and thus, a low evaporative demand. This was not the case for broadleaf species whose larger leaf area requires greater evaporative demand which incurs a tradeoff between maximizing carbon gain and minimizing water loss when exposed to drought conditions. Contrary to their findings, loblolly seedlings in our study suffered nearly 50 percent mortality at both sites. We are uncertain why observed loblolly mortality was so high in this study.

A general trend was observed that smaller seedlings had better survivorship than larger seedlings. This trend was consistent for all species at both sites except for the longleaf at S1. The difference in seedling sizes between the mortality and surviving group was statistically significant only for chinkapin and persimmon seedlings at S2. Unfortunately the only seedlings available from nurseries at the time of planting varied in

sizes within species. Many of the chinkapin oak seedlings were large and cumbersome to plant at S2 which could explain why the mortality rate was so high based on the findings of Casperson and Kobe (2001). This project could have potentially been improved by sorting through the seedlings and creating a more uniform size class. Additionally, this project could have been more informative if the study duration included two or three growing seasons in order to gain a more accurate idea of how wild hog predation of planted seedlings changes over time.

4.4.8. Implications of Results and Management Suggestions

In summary, we suggest the level of damage among planted seedlings is likely driven by a wild hog's preference and familiarity with the species as a food source. Reducing the amount of time wild hogs spend in the plantation should lower the likelihood of the animals becoming educated about seedlings as a preferred food source. Not all seedlings species are preferred and wild hog presence in the area does not guarantee seedlings will be targeted. A combination of ecological factors discussed earlier, rather than just a single factor, is likely to influence the severity of wild hog damage in forest plantations.

The results of this project are an additional step towards better understanding variations of wild hog behavior in young forest plantations. Research projects like these are important for finding how changes in management could improve seedling survival during the stage of vulnerability to wild hogs. The following recommendations are based on the knowledge gained through observations made during this study and will require further testing as to their effectiveness. The recommendations are as follows: 1) request

logging crews to pile woody debris away from the site where seedlings will be planted rather than scattering the debris, 2) wild hog population reduction through hunting and trapping should be implemented during the first three growing seasons in young forest plantations once seedlings are observed to be damaged, 3) add seedling tubes adequately anchored in the soil to make it more difficult for hogs to access seedlings, and 4) delay herbaceous release treatments by a year to allow vegetation to grow up around the seedlings making them more difficult for wild hogs to find or access.

CHAPTER 5

Conclusion

It would appear efforts to raise awareness about the importance the longleaf pine (*Pinus palustris* Mill.), from both a historic and ecological standpoint, are starting to come to fruition. Data from the Forest Inventory and Analysis (FIA) program indicated longleaf had been declining in the southern United States (US) from 1970-2010 (Oswalt et al. 2012), but in recent years it seems this trend has been reversed (Guldin et al. 2016). The 2015 south-wide estimate for longleaf acreage was 4.7 million (LPC 2017), this represents a 30 percent increase from 2010 (Guldin et al. 2016). The recent increase in longleaf acreage can primarily be attributed to the efforts of America's Longleaf Restoration Initiative (ARLI). Founded in 2009, the ARLI involved federal resources to recover longleaf across its former range. The overall goal of ARLI is to restore longleaf to 8 million acres by 2025 (LPC 2017). More than half of the restoration work accomplished in 2016 occurred in Florida and Alabama. According to progress reports from the Longleaf Pine Initiative (LLPI) (an initiative created by the Natural Resources Conservation Service to support the ARLI), Alabama holds the annual record for the most acres put into longleaf restoration (progress reports and other information about the LLPI is available online at <https://www.nrcs.usda.gov/wps/portal/nrcs/detailfull/national/>

programs/initiatives/?cid=nrcsdev11_023913). Alabama contains a large number of counties given high restoration priority because of their value in connecting existing longleaf landscapes. Due to the high number of priority counties in Alabama, the state is critical to longleaf restoration efforts.

The majority of longleaf establishment since 2000 has taken place in the eastern portion of the Coastal Plains region (Oswalt et al. 2012). Among the recent planting efforts in 2016, 79 and 21 percent were on private and public lands, respectively (LPC 2017). It is apparent private landowners have an important role in the successful reestablishment of longleaf across its native range. In Alabama, non-industrial private landowners hold 88 percent of the forest land (AFC 2016); therefore, insuring the successful establishment of longleaf on non-industrial private lands should be a priority.

The wild hog (*Sus scrofa*) has shown itself capable of hindering longleaf restoration efforts; this animal's propensity for damaging longleaf seedlings has been a topic frequently discussed in previous chapters. Wild hogs can have a substantial economic impact on young longleaf plantations of non-industrial private landowners, most of whom must personally absorb the financial costs of wild hog damage and control. In most cases, financial assistance is not available to help compensate for wild hog damage or to offset control costs. In contrast to non-industrial private lands, the impact of damage to longleaf established on public or private industrial lands is likely minimal due to the funds and resources available to recover any losses from wild hogs.

As more acres of longleaf are planted across the South in years to come, it is likely the corresponding damage from wild hogs will increase as well. Historical reports

showed this to be the case nearly 70 years ago when entire longleaf plantations in Alabama, Florida, Louisiana, and Mississippi resulted in crop failure due to wild hogs after intensive planting efforts by the US Forest Service (Lucas 1977). The future damage to longleaf plantations may be even more extensive than in the 1950's as recent wild hog populations have grown more abundant. Repeated years of hog-damage to a non-industrial private landowner's plantation may prove enough to discourage them away from producing longleaf. With non-industrial private landowners being a critical component of longleaf restoration efforts, losing the support of this group would prove detrimental in reaching the goal of 8 million acres by 2025.

Management suggestions are made in the third and fourth chapters in an effort to help forest plantation owners plan for wild hog damage and reduce seedling mortality. The impact of wild hogs can be remediated by financially preparing for the possibility of damage. The economic analyses from chapter 3 showed how replanting the year after wild hog damage occurs can reduce most of the cost of damage. Even postponing replanting an extra year allows the majority of stand profitability to be recovered compared to not replanting. When simulations were run in a scenario where replanting does not occur, the cost of damage quickly exceeded profitability. The sensitivity analyses tables created for this project are applicable for landowners or resource professionals utilizing a shelterwood system to grow longleaf on a 60 year rotation. Applying the sensitivity analyses tables to scenarios outside of the parameters they were intended could have misleading results; however, they are useful in getting people thinking about planning for wild hog damage and control before trees have even been

planted. This project serves as a starting point in creating awareness of the economic ramifications and scale of wild hog damage to forest plantations.

In chapter 4 we showed tree species and ecological factors can influence the amount of damage to forest plantations from wild hogs. Planting highly preferred tree species in an area where there are wild hogs does not guarantee direct mortality will occur. Results from the field study suggest seedling depredation in forest plantations is largely a learned behavior among wild hogs. In populations where wild hogs were naïve to eating planted seedlings, plantation damage was minimal. Even so, it is important for individuals interested in establishing forest plantations to be mindful of the ecological conditions in and around the plantation. If planting a highly preferred tree species in an area commonly used by wild hogs, it is important to be proactive in discouraging wild hogs from discovering seedlings. This can be accomplished in a number of ways but a combination of methods might produce the best outcome for protecting seedlings during the first three growing seasons. We recommend the active management of wild hogs for at least the first three growing seasons until seedlings reach a size where they are no longer desirable as a food source. This does not guarantee wild hog damage will not occur after this point but the scale of damage will likely not be as large as in the first three growing seasons.

Based on field observations it appears there is merit to using tree tubes to discourage wild hogs away from seedlings and towards more easily obtained food sources. Tree-tube stakes should be adequately anchored in the ground to reduce the likelihood of damage to seedlings by rooting. In conjunction with tree tubes, reducing

desirable food sources in the plantation will also help in limiting the possibility of wild hogs discovering seedlings. In situations where cutover sites have been planted it could prove beneficial to remove woody debris; doing so would give wild hogs one less reason to enter the plantation area in search of abundant sources of invertebrates. There is a positive association between rooting frequency by wild hogs and the presence of coarse woody debris (Zengel and Conner 2008). Additionally, hunting over the plantation in the spring, when seedling damage is most likely to occur, could deter wild hogs from the area until vegetation has grown thick enough to help protect seedlings. Delaying the herbaceous release treatment will also allow for thicker layers of vegetation making it more difficult to find and access planted seedlings.

Hog density is also an important factor influencing the severity of damage in forest plantations; having more wild hogs in an area increases the chance of plantation damage occurring. Survey results indicated plantation owners in the Coastal Plains region of Alabama experienced higher levels of damage than those in northern regions where wild hog densities are lower. Population reduction on the property can be an important management tool for protecting forestry investments like young pine stands. In general, survey respondents did not feel the current control methods available for wild hog removal were very efficient. However, survey respondents did associate the highest efficiency ratings with hunting with dogs, yet this method was not as commonly utilized compared to other control methods. Some speculate hunting with dogs can cause home range shifts and harass wild hogs away from particular areas of concern (Engeman et al. 2007, Gaston 2008); therefore, this could be a very useful management technique for plantation owners needing to protect their seedlings.

Wild hog research pertaining to forestry has become substantially less common as other more popular topics capture public attention. Over the past decade wild hog research has been dominated by topics related to diseases, parasites, and genetics. The two subject categories, diseases/parasites and genetics, account for 40 percent of key words used to describe entries in the bibliography created for this project (Chapter 2). This study is one step toward understanding the economic impact wild hogs have on the forest industry and the ecological factors influencing forest plantation damage. Future research is needed to examine the possibility of insurance policies to protect young forest stands against wild hog damage. The need for such policies will become apparent as more longleaf plantations are impacted by growing and expanding hog populations. The management suggestions mentioned previously warrant more research to test their efficiency in minimizing seedling damage in forest plantations. Similar experiments as the study performed in chapter 4 can be repeated over extended growing periods to gain a better understanding of why seedling depredation varies based on location.

In conclusion, the long-term nature of forestry investments warrants the threat from wild hogs be taken seriously. Crop failure in a young forest plantation can offset a rotation by years and carries considerably more economic loss into the future compared to traditional/annual agricultural crops. The costs of wild hog introduction into an area may be felt over many years as their presence would create uncertainty and risk to timber investments. The short-term protection and minimization of seedling depredation in young forest plantations is the most realistic solution to reducing the impact of wild hogs on forestry and timber resources.

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Appendix A

FERAL SWINE SURVEY – 2016



Auburn University School
of Forestry and Wildlife
Sciences



Alabama Cooperative
Extension System

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Alabama Cooperative Extension System

Please make corrections to name, address and ZIP Code, if necessary.

The information you provide will be used for statistical purposes only. In accordance with the Confidential Information Protection provisions of Title V, Subtitle A, Public Law 107-347 and other applicable Federal laws, your responses will be kept confidential and will not be disclosed in identifiable form to anyone other than employees or agents. By law, every employee and agent has taken an oath and is subject to a jail term, a fine, or both if he or she willfully discloses ANY identifiable information about you or your operation. Response is **voluntary**.

The time required to complete this information collection is estimated to average 15 minutes per response, including the time for reviewing instructions, searching existing data sources, gathering and maintaining the data needed, and completing and reviewing the collection of information.

SECTION 1 – GENERAL

For the purposes of this study, wild pigs refer to all species of feral swine, feral hogs, and wild boar.

1. On June 1, 2015, how many forest acres did you:

- a. Own?
- b. Rent or Lease from others or use Rent Free?
(**Exclude** land used on an animal unit month (AUM) basis, BLM and Forest Service land.)
- c. Rent to others?

Acres
100
101
102

2. In what state and county is the majority of your forest land located?

State

County

OFFICE USE
103
104

3. During the last three years, have wild pigs been present in the county reported in item 2?

105 ☐ Yes ☐ No ☐ Don't Know

4. During the last three years, have wild pigs been present on your property?

106 ☐ Yes ☐ No ☐ Don't Know

SECTION 2 – TIMBER STANDS

(Please answer the following in terms of the forest acres specified in question 1 of section 1)

1. Were any tree seedlings planted on your forest land in 2013-2015?

107 ☐ Yes – Continue ☐ No – Go to Item 3

2. Please report acreage planted in 2013-2015.

	Unit	2013 Stand	2014 Stand	2015 Stand
a. What tree species were planted? (List top 2 dominant species planted)	Species	108	109	110
(i) Were the seedlings planted containerized?	Check Mark	111 <input type="checkbox"/> Yes <input type="checkbox"/> No	112 <input type="checkbox"/> Yes <input type="checkbox"/> No	113 <input type="checkbox"/> Yes <input type="checkbox"/> No
(ii) What was the total payment spent prepping the site before planting?	Dollars	114 \$	115 \$	116 \$
(iii) How many acres of each tree species were planted?	Acres	117	118	119
b. How many trees per acre were planted?	Trees Per Acre	120	121	122
c. Around how many trees per acre survived 1 year after planting?	Trees Per Acre	123	124	125
(i) How many total acres of this stand were damaged by wildlife in general?	Acres	126	127	128
d. Were wild pigs present in any part of the stand? (If Yes, continue. If No or Don't Know, go to Item 3.)	Check Mark	129 <input type="checkbox"/> Yes <input type="checkbox"/> No <input type="checkbox"/> Don't Know	130 <input type="checkbox"/> Yes <input type="checkbox"/> No <input type="checkbox"/> Don't Know	131 <input type="checkbox"/> Yes <input type="checkbox"/> No <input type="checkbox"/> Don't Know
(i) Was this stand damaged by wild pigs? (If Yes, continue. If No or Don't Know, go to Item 3.)	Check Mark	132 <input type="checkbox"/> Yes <input type="checkbox"/> No <input type="checkbox"/> Don't Know	133 <input type="checkbox"/> Yes <input type="checkbox"/> No <input type="checkbox"/> Don't Know	134 <input type="checkbox"/> Yes <input type="checkbox"/> No <input type="checkbox"/> Don't Know
(ii) How many acres of this stand were damaged by wild pigs?	Acres	135	136	137
(iii) On the acres that were damaged by wild pigs, what percentage of seedlings were destroyed?	Percentage	138	139	140

3. On your property, in which type of habitat did you observe **wild hog activity** between 2013-2015? (Check one for each)

		FREQUENCY OF SIGHTING		
		Never	Rarely	Frequently
a. Bottomland hardwood	141	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>
b. Hardwood plantation	142	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>
c. Upland hardwood	143	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>
d. Agricultural field	144	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>
e. Grassland/pastureland	145	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>
f. Pine plantation	146	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>
g. Clearcut	147	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>
h. Developed area/roads	148	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>
e. Swamp	149	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>
f. Mixed pine & hardwood	150	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>
g. Natural pine	151	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>
h. Other (Specify: _____)	152	<input type="checkbox"/>	<input type="checkbox"/>	<input type="checkbox"/>

4. What percent of each forest type would best describe your forest land? (*Total should equal 100%*)

- 153 _____ % Bottomland hardwood
 154 _____ % Hardwood plantation
 155 _____ % Upland hardwood
 156 _____ % Agricultural field
 157 _____ % Grassland/pastureland
 158 _____ % Pine plantation
 159 _____ % Clearcut
 160 _____ % Developed area/roads
 161 _____ % Swamp
 162 _____ % Mixed pine and hardwood
 163 _____ % Natural pine
 164 _____ % Other → Please specify _____

SECTION 3 – CONTROL

(Please answer the following in terms of the forest acres specified in question 1 of section 1)

1. In 2015, were any control methods used on your operation in order to **reduce** or **prevent** damage from wild pigs?

- 165
 1 ☐ Yes – Continue 3 ☐ No – Go to Item 3

2. In 2015, which of the following control methods were used on your operation in order to **reduce** or **prevent** damage from wild pigs? If the method was used in 2015, how much did that method cost (including labor) and how effective was it?

Control Method	Used on operation in 2015?	Cost (Include labor) (Dollars)	How effective was the control method? Check one per row			
			Not Effective	Slightly Effective	Moderately Effective	Very Effective
a. Shooting wild pigs on sight	166 1 <input type="checkbox"/> Yes 3 <input type="checkbox"/> No	167 \$	168 1 <input type="checkbox"/>	2 <input type="checkbox"/>	3 <input type="checkbox"/>	4 <input type="checkbox"/>
b. Hunting wild pigs with dogs	169 1 <input type="checkbox"/> Yes 3 <input type="checkbox"/> No	170 \$	171 1 <input type="checkbox"/>	2 <input type="checkbox"/>	3 <input type="checkbox"/>	4 <input type="checkbox"/>
c. Hunting wild pigs without dogs	172 1 <input type="checkbox"/> Yes 3 <input type="checkbox"/> No	173 \$	174 1 <input type="checkbox"/>	2 <input type="checkbox"/>	3 <input type="checkbox"/>	4 <input type="checkbox"/>
d. Aerial hunting (helicopter)	175 1 <input type="checkbox"/> Yes 3 <input type="checkbox"/> No	176 \$	177 1 <input type="checkbox"/>	2 <input type="checkbox"/>	3 <input type="checkbox"/>	4 <input type="checkbox"/>
e. Trapping and removing wild pigs	178 1 <input type="checkbox"/> Yes 3 <input type="checkbox"/> No	179 \$	180 1 <input type="checkbox"/>	2 <input type="checkbox"/>	3 <input type="checkbox"/>	4 <input type="checkbox"/>
f. Repellents for wild pigs	181 1 <input type="checkbox"/> Yes 3 <input type="checkbox"/> No	182 \$	183 1 <input type="checkbox"/>	2 <input type="checkbox"/>	3 <input type="checkbox"/>	4 <input type="checkbox"/>
g. Other (Specify: _____)	184 1 <input type="checkbox"/> Yes 3 <input type="checkbox"/> No	185 \$	186 1 <input type="checkbox"/>	2 <input type="checkbox"/>	3 <input type="checkbox"/>	4 <input type="checkbox"/>

3. In 2015, did you use **electric** fencing on your operation **specifically** to reduce damage by wild pigs?

- 187
 1 ☐ Yes – Continue 3 ☐ No – Go to Item 4

a. How much did it cost to install this **electric** fencing?

Dollars
188

b. At the time of installation, how many years was the expected useful life of this **electric** fencing? . .

Years
189

SECTION 3 – CONTROL

(Please answer the following in terms of the forest acres specified in question 1 of section 1)

4. In 2015, did you use **non-electric** fencing on your operation **specifically** to reduce damage by wild pigs?

190

☐ Yes – Continue ☐ No – Go to Item 5

- a. How much did it cost to install this **non-electric** fencing?

Dollars

191

- b. At the time of installation, how many years was the expected useful life of this **non-electric** fencing?

Years

192

None

Head

193

5. In 2015, how many wild pigs were **killed** on your land?

☐

6. In 2015, did you seek help from a county, state, or federal agency because of the damage due to wild pigs on your land?

194

☐ Yes

☐ No

☐ NA – No damage by wild pigs

7. Please choose **only one** of the statements below that best describes your opinion:

I **have** some wild pigs on my land, and I wouldn't change how many. 195 ____ (go to **1A** below)

I **have** some wild pigs on my land, but would prefer to have more. 196 ____ (go to **1A** below)

I **have** some wild pigs on my land, but would prefer to have fewer. 197 ____ (go to **1B** below)

I **have** some wild pigs on my land, but would prefer to have none. 198 ____ (go to **1B** below)

I **don't have** wild pigs on my land, and I would prefer to keep it that way. 199 ____ (go to **1C** below)

I **don't have** wild pigs on my land, but would prefer to have them. 200 ____ (go to **1D** below)

Please answer **only one** of the following questions based on your response above.

- 201 **1A.** What is the minimum you would have to be paid per acre to allow permanent eradication of wild pigs from ALL of your land? (Check **ONE** dollar amount below)

- 202 **1B.** If eradication costs were spread across all farmers in your region, what is the most you would pay to guarantee permanent eradication of wild pigs from your land? (Check **ONE** dollar amount below)

- 203 **1C.** What would be the most you would pay to guarantee that wild pigs never spread to your land? (Check **ONE** dollar amount below)

- 204 **1D.** What is the minimum you would have to be paid per acre to never have wild pigs on your land? (Check **ONE** dollar amount below)

1 ____\$0

4 ____\$5/acre per year

7 ____\$10/acre per year

10 ____\$15/acre per year

2 ____\$20/acre per year

5 ____\$25/acre per year

8 ____\$50/acre per year

11 ____\$75/acre per year

3 ____\$100/acre per year

6 ____\$200/acre per year

9 ____\$500/acre per year

12 ____\$1,000/acre per year

SECTION 4 – HUNTING

(Please answer the following in terms of the forest acres specified in question 1 of section 1)

1. In 2015, did you or anyone else hunt on your land?

205

☐ Yes – Continue

☐ No – Go to Section 5

☐ Don't Know – Go to Section 5

2. Please answer the following items regarding hunting of wildlife on your operation in 2015. Check all that apply

Hunting	Wild Pigs	Deer	Upland Birds	Waterfowl	Turkeys	Other
a. Which animals did you or your immediate family hunt on your land in 2015?	206 <input type="checkbox"/>	207 <input type="checkbox"/>	208 <input type="checkbox"/>	209 <input type="checkbox"/>	210 <input type="checkbox"/>	211 <input type="checkbox"/>
b. Which animals did anyone other than your immediate family hunt on your land in 2015 without paying?	212 <input type="checkbox"/>	213 <input type="checkbox"/>	214 <input type="checkbox"/>	215 <input type="checkbox"/>	216 <input type="checkbox"/>	217 <input type="checkbox"/>
c. Which animals did anyone pay to be allowed to hunt on your land in 2015? (Exclude guide or outfitting services)	218 <input type="checkbox"/>	219 <input type="checkbox"/>	220 <input type="checkbox"/>	221 <input type="checkbox"/>	222 <input type="checkbox"/>	223 <input type="checkbox"/>
d. For which animals did you provide a guide or outfitting service to paying hunters on your land in 2015?	224 <input type="checkbox"/>	225 <input type="checkbox"/>	226 <input type="checkbox"/>	227 <input type="checkbox"/>	228 <input type="checkbox"/>	229 <input type="checkbox"/>
e. Which animals have experienced a decline in population on your land in the last few years due to wild pigs?		230 <input type="checkbox"/>	231 <input type="checkbox"/>	232 <input type="checkbox"/>	233 <input type="checkbox"/>	234 <input type="checkbox"/>

3. During 2015, what was your net income from:

None

a. **Wild pig hunting** on your land?

☐

b. **All hunting activities** on your land?

☐

Dollars
235
236

SECTION 5 – DEMOGRAPHICS

1. What was your age on December 31, 2015?

238

2. What is your gender?

☐ Male ☐ Female

4. Do you make the majority of the decisions about your forest land?

239

☐ Yes ☐ No

5. Do you have a hired manager for your forest?

240

☐ Yes ☐ No

6. Do you currently live on your forest?

241

☐ Yes ☐ No

7. In what year did you begin to make day-to-day decisions for your forest land?

Year (YYYY)
242

8. In 2015, what was your total household income before taxes? Check one only

- 243
- 1 ☐ Less than \$1,000
 - 2 ☐ \$1,000 to \$2,499
 - 3 ☐ \$2,500 to \$4,999
 - 4 ☐ \$5,000 to \$9,999
 - 5 ☐ \$10,000 to \$24,999
 - 6 ☐ \$25,000 to \$49,999
 - 7 ☐ \$50,000 to \$99,999
 - 8 ☐ \$100,000 to \$249,999
 - 9 ☐ \$250,000 to \$499,999
 - 10 ☐ \$500,000 to \$999,999
 - 11 ☐ \$1,000,000 or more

9. In 2015, what percent of your total household income came from forest related activities? Check one only

- 244
- 1 ☐ None
 - 2 ☐ 1% to 25%
 - 3 ☐ 26% to 50%
 - 4 ☐ 51% to 75%
 - 5 ☐ 76% to 99%
 - 6 ☐ 100%